Molecular modeling to elucidate the dynamic interaction process and aggregation mechanism between natural organic matters and nanoplastics

Chi Zhang, Zhiyu Zhou, Mengning Xi, Haozhe Ma, Junhao Qin, Hanzhong Jia

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4	Chi Zhang <sup>a,b,1</sup> , Zhiyu Zhou <sup>a,1</sup> , Mengning Xi <sup>a</sup> , Haozhe Ma <sup>a</sup> , Junhao Qin <sup>c,*</sup> ,
5	Hanzhong Jia <sup>a,b</sup> ,*
6	<sup>a</sup> College of Natural Resources and Environment, Northwest A & F University, Xianyang 712100,
7	China
8	<sup>b</sup> Key Laboratory of Low-Carbon Green Agriculture in Northwestern China, Ministry of
9	Agriculture and Rural Affairs, Xianyang 712100, China
10	° Guangdong Laboratory for Lingnan Modern Agriculture, Guangdong Provincial Key Laboratory
11	of Agricultural & Rural Pollution Abatement and Environmental Safety, College of Natural
12	Resources and Environment, South China Agricultural University, Guangzhou 510642, China
13	
14	
15	<sup>1</sup> Contributed equally to this work
16	*Corresponding authors
17	E-mails: J_Qin@scau.edu.cn (J. Qin); jiahz@nwafu.edu.cn (H. Jia)
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## 28 Abstract

29 The interactions of nanoplastics (NPs) with natural organic materials (NOMs) 30 dominate the environmental fate of both substances and the organic carbon cycle. Their 31 binding and aggregation mechanisms at the molecular level remain elusive due to the 32 high structural complexity of NOMs and aged NPs. Molecular modeling was used to 33 understand the detailed dynamic interaction mechanism between NOMs and NPs. 34 Advanced humic acid models were used, and three types of NPs, i.e., polyethylene (PE), polyvinyl chloride (PVC), and polystyrene (PS), were investigated. Molecular 35 dynamics (MD) simulations revealed the geometrical change of the spontaneous 36 37 formation of NOMs-NPs supramolecular assemblies. The results showed that pristine 38 NPs initially tend to aggregate homogeneously due to their hydrophobic nature, and 39 then NOM fragments are bound to the formed NP aggregates mainly by vdW interaction. 40 Homo- and hetero-aggregation between NOMs and aged NPs occur simultaneously 41 through various mechanisms, including intermolecular forces and Ca<sup>2+</sup> bridging effect, 42 eventually resulting in a mixture of supramolecular structures. Density functional theory calculations were employed to characterize the surface properties and reactivity 43 44 of the NP monomers. The molecular polarity indices for unaged PE, PS, and PVC were 45 3.1, 8.5, and 22.2 kcal/mol, respectively, which increased to 43.2, 51.6, and 42.2 46 kcal/mol for aged NPs, respectively, indicating the increase in polarity after aging. The 47 vdW and electrostatic potentials of NP monomers were visualized. These results clarified the fundamental aggregation processes, and mechanisms between NPs and 48 49 NOMs, providing a complete molecular picture of the interactions of nanoparticles in 50 the natural aquatic environment.

51 Keywords: Natural organic matters; Aged nanoplastics; Interaction processes;
52 Aggregation mechanism; Molecular modeling

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### 54 1. Introduction

55 Natural organic matters (NOMs) play a critical role in a variety of 56 (bio)geochemical processes in natural water environments by directly affecting 57 microbial activity, co-aggregation, contaminant retention, water quality, and element 58 cycling. The environmental behaviors of NOMs are also closely related to the global 59 carbon cycle and climate change by preserving and converting organic carbon [1, 2]. 60 Despite their importance, a thorough and precise understanding of NOMs' composition 61 and structure is still lacking. Structurally, NOM is extremely complex and 62 heterogeneous as a supramolecular mixture of chemically distinct organic compounds that are abundant in chemically diverse functional groups and active sites. The major 63 64 functional groups in NOM have long been clearly identified, such as carboxylic, phenolic, alcoholic, carbonyl, amino, etc. The compositional complexity and structural 65 diversity of NOMs present a severe challenge to experimental characterization and 66 67 building of realistic models, but nowadays computational chemistry methods have 68 made this problem tractable [3-5].

69 In nature, NOMs interact closely with organic pollutants through mechanisms such 70 as adsorption, co-precipitation, and homogeneous/heterogeneous aggregation through 71 different mechanisms such as  $\pi$ - $\pi$  interaction, hydrophobic interaction, H-bonding, 72 electrostatic attraction, steric hindrance effect, and so forth [6], thus governing the 73 environmental fate of pollutants and profoundly impacting the distribution and 74 preservation of organic carbon. Nanoplastics (NPs), as a class of emerging 75 contaminants, have a particle size between 1 nm and 100 nm [7] and are generated from 76 plastic production by natural or anthropogenic factors, such as photooxidation and 77 microbial degradation [8, 9]. Due to their chemical stability, persistence, and 78 bioaccumulation, the environmental behaviors of NPs, in particular their interactions 79 with different environmental pollutants during their transport through the marine or 80 sedimentary environment into ecosystems, have received considerable research 81 attention in recent years [10-13]. Compared to microplastics (MPs, <5mm), the

82 molecular-level properties and environmental implications of NPs in natural systems 83 have been relatively less studied because of methodological challenges. For instance, it 84 is still technically challenging to extract NPs from natural environments, and their 85 surface properties are not clearly identified [14-17]. There is also research suggesting 86 that NPs are more abundant and hazardous in the environment than MPs [18]. The 87 transport and fate of NPs are greatly determined by interactions with environmental media, such as aggregation with NOMs and deposition onto minerals' surfaces [19-22]. 88 89 Therefore, it is imperative to understand the environmental processes and behaviors of 90 NPs with environmental media, thus providing significant implications for their 91 ecological risk assessment.

92 The interactions between NOMs and NPs are of fundamental importance and largely control the environmental fate of both materials, as well as the stabilization and 93 94 cycling of organic carbon in nature [23]. NOMs can have an impact on the transport 95 and distribution of NPs in the aquatic environment, because the presence of NOMs can alter the surface properties of NPs, affecting their aggregation and mobility [24]. NOM 96 97 can also be adsorbed onto the surface of NPs, changing their chemical reactivity and 98 affecting the adsorption of other contaminants, such as metals and organic pollutants, 99 onto the NP surface [25]. Furthermore, the evaluation of the effect of NOMs on the 100 behaviors of NPs was the first critical step in assessing the potential environmental 101 hazards of NPs in ecosystems. However, a detailed molecular-level picture of the 102 dynamic interaction processes and the underlying mechanism between NOMs and NPs 103 is still elusive. It is difficult to directly observe and decipher the complicated NOM–NP 104 aggregating behaviors only relying on traditional experimental techniques. There is an 105 urgent need for a better understanding of such processes that can provide more 106 quantitative information, such as the determination of fundamental mechanisms 107 (homo-aggregation or hetero-aggregation), identification of active sites, observation of dynamic aggregation processes, and characterization of binding and aggregating 108 109 structures.

110 With the rapid development of theoretical and computational chemistry and 111 computer hardware, computational simulations have been proven an essential and 112 powerful approach in quantitatively revealing aqueous environmental processes and 113 reactions at the molecular scale beyond the conventional resolution of analytical 114 methods, based on the identification and analysis of a range of physicochemical 115 properties, such as geometrical characteristics, thermodynamics, electronic structures, 116 and dynamics of substances [5, 26-33]. For example, Xue et al. [34] reviewed the 117 application of computational modeling methods at various scales in understanding the 118 interface behavior of pollutants, especially in aquatic environments. Li et al. [31] 119 reviewed recent progress in using environmental theoretical calculations to elucidate 120 the transformation mechanisms of contaminants. In addition, this group has extensively 121 revisited the application of density functional theory (DFT) calculations to the prediction of active sites and reaction selectivity, as well as oxidant activation 122 123 mechanisms in the field of advanced oxidation processes (AOPs) [30]. However, 124 research providing deep insights into the interactions between NOMs and NPs has 125 rarely been conducted using theoretical calculation methods. In a recent study, the 126 dynamic properties and microscopic aggregation structures related to the NPs-NOMs 127 assemblage were studied by molecular dynamics (MD) simulations, which acquired 128 deep insights into the interparticle interactions between both substances [35]. However, this study only incorporated virgin plastics, and the aged forms of NPs were not 129 130 considered. According to previous research [23,36-38], oxygen-containing functional 131 groups, including carboxyl, phenolic, and carbonyl groups, are generated after aging, 132 which leads to the modification of the surface morphology of MPs, such as specific 133 surface area and roughness. Therefore, it can be expected that different surface 134 properties of aged NPs would lead to more significant and interesting phenomena and 135 aggregation mechanisms with NOMs in comparison to pristine NPs, given that NOMs 136 have a high degree of structural complexity and heterogeneity, the chemical composition of which is very diverse, and aged NPs also have a wide variety of O-137

138 containing functional groups.

139 In the present study, by performing MD simulations and DFT calculations, we 140 explored the dynamic interactions and elucidated the detailed molecular-level 141 mechanism between NOMs and three types of NPs, polyethylene (PE), polyvinyl 142 chloride (PVC), and polystyrene (PS) in both pristine and aged form. The dynamic 143 binding and aggregation processes between two substances are qualitatively revealed 144 by the movement trajectories and statistical data. Analysis of the interaction energies between NOM and NP fragments and characterization of the direct intermolecular 145 binding and local coordination geometries of  $Ca^{2+}$  bridging provides a profound insight 146 147 into the aggregation mechanism. DFT calculations allow a quantitative analysis of the 148 reactivity and active sites of NPs, and weak interactions between NOM and PS 149 fragments are presented visually. These results establish a fundamental basis for the 150 understanding of the fate of NPs in the aquatic environment and the stabilization of 151 organic C.

## 152 2. Methodology

## 153 **2.1. Computational models**

154 **2.1.1.** Leonardite humic acid

155 Vienna Soil Organic Matters Modeler 2 (VSOMM2) is an online tool to generate theoretical models of humic substances [39, 40] based on the data derived from 156 157 standardized samples of the International Humic Substances Society (IHSS). The 158 NOMs condensed phase models constructed by VSOMM2 fundamentally rely on the 159 concept of building blocks (BBs), which represent the simple units in organic 160 components, and the implementation of a genetic algorithm randomly results in the 161 chemical and structural diversity of the final models. On this website, the primary elemental composition and organic carbon fractions are required as input, and the 162 modeler calculates and generates a diverse combination of heterogeneously distributed 163

164 NOM fragments automatically that match the input.

165 In this study, VSOMM2 was used to build theoretical models of the supramolecular structure of Leonardite humic acid (LHA) according to the elemental and organic 166 167 composition of the IHSS. LHA has been extensively used in the molecular modeling of 168 NOMs in previous studies [5, 27]. The created LHA model was composed of 100 169 building blocks with 10 building blocks per molecule. The detailed element and organic 170 composition of the generated LHA are listed in Table S1. The molecular geometries of 171 the 10 deprotonated NOM fragments are illustrated in Fig. S1. These molecules 172 collectively constitute the NOMs' supramolecular association.

## 173 **2.1.2.** Nanoplastics

174 Both pristine and aged NPs (Fig. S2) were investigated in our simulations with the 175 purpose of clarifying the effect of oxygen-containing functional groups on aggregation 176 behaviors. Each monomer of the PE and PVC chain had a degree of polymerization of 177 10, and each PS monomeric chain contained 5 polymerized units. There were 16 such 178 NP monomers in each simulation system. For pristine NPs, the models we used were 179 similar to those in previous studies. For aged NPs, due to their structural complexity, 180 the models have not yet been developed, so we had to use the simplified models. Simplified models of aged NPs were constructed by incorporating O-containing 181 182 functional groups according to previous studies [35, 41]. In brief, each monomer 183 contained four carbonyls, two hydroxyls, and one carboxyl. The carboxylic groups of 184 NOM and NP fragments were deprotonated (in a circumneutral environment) due to the 185 relatively low  $pK_a$  of these groups. Each monomeric chain of aged NPs possessed a net 186 charge of -1, and the 10 NOM fragments had different negative charges depending on 187 the number of carboxylate groups.

## 188 2.1.3. Simulation systems

NOMs–NPs coexisting binary systems (Fig. 1) were constructed. NOMs and NPs
 molecules were initially placed at random positions within the simulation cell. This

191 setup was designed to enable the direct observation of the spontaneous aggregation 192 processes of both materials. Alternatively, one can first build a homo-aggregate of NP 193 nanoparticles, then possibly age its surface, and then insert these NOM molecules into 194 the solution. This setup should be more reasonable but also more complicated and 195 therefore not adopted in this work. Moreover, to clearly reveal the effect of their 196 interactions on the behaviors of both NOMs and NPs, unary systems that contained 197 only NOMs or NPs were also simulated. There were 13 individual simulation systems in all: 6 binary systems and 7 unitary systems. For each system, NOM/NP fragments 198 199 were first randomly distributed in the simulation box (approximately 76.5Å  $\times$  76.5Å  $\times$ 200 76.5Å), and then salt ions and solvent water molecules were randomly placed in the region to simulate the solution environment. It is well known that Ca<sup>2+</sup> and Na<sup>+</sup> are the 201 most common salt ions. Therefore, Ca<sup>2+</sup> ions were added to counterbalance the net 202 negative charges from NOMs and aged NPs, and additionally ~0.1 mol/L NaCl was 203 204 added in each system. The detailed compositions of all the systems are summarized in 205 Table S2. All of the simulation cells were periodically repeated.

206

Fig. 1. Computational models of original systems used for MD simulations. NOMs and
NPs fragments are shown as sticks; salt ions are represented by balls. Color code: N =
blue; H = white; O = red; S = yellow; Na = green; Ca = purple; Cl = orange. The C in
NOMs and NPs are colored by cyan and brown, respectively. The water molecules are
not shown for clarity.

212

### 213 **2.2 MD setup**

All MD simulations were carried out by GROMACS 2018.4 [42]. Amber 03 force field [43] was used to describe interatomic interactions. A cutoff value of 12 Å was used for short-range electrostatic and van der Waals (vdW) interactions. The PME summation method [44] was employed for the summation of long-range electrostatic interactions. The SPC/E water model was used [45]. The rigidity of water was constrained by the LINCS algorithm [46]. The atomic charges were derived by DFT 220 calculations (see the details in *Section 2.3*).

The propagation of Newtonian equations of motion was realized via the leap-frog algorithm. Berendsen barostat [47] and velocity-rescale thermostat [48] were employed to maintain the pressure and temperature, respectively. For each simulation, we first performed a pre-equilibration for 0.5 ns in the NVT ensemble at 298 K and then 1–2 ns equilibration in the NPT ensemble at 1 atm and 298 K. Subsequently, a production run for 100 ns with a timestep of 2 fs in NVT ensemble was carried out. Data of the last 20 ns were used for statistics.

## 228 **2.3 DFT calculations for surface reactivity**

DFT calculations were used to give a more comprehensive understanding of the 229 230 interaction mechanism at atomic and electronic scales. All of the quantum chemistry 231 calculations were conducted by the Gaussian 16 program. The geometry optimizations 232 of the 6 monomeric chains of NPs (i.e., pristine and aged PE, PVC, and PS) and 10 233 individual NOM fragments were performed at the quantum chemistry level of def2-234 TZVP basis set [49] and B3LYP functional [50, 51], complemented by DFT-D3 235 dispersion correction [52]. The partial atomic charges of all the NOMs and NP 236 monomers were derived by the Multifwn 3.8 program [53]. The acquired RESP2 (0.5) 237 atomic charges [54] were used in MD simulations.

238 Several key descriptors were analyzed on the basis of the structures and 239 wavefunctions by Multifwn 3.8 [53], including electrostatic potential (ESP), vdW 240 potential, and molecular polarity index (MPI) of NP monomers. ESP is quite common 241 and useful for intuitively disclosing possible electrostatic interaction of the investigated 242 chemical entity with the external environment, and ESP is usually mapped on an 243 electron density isosurface of 0.001 a.u. [55]. Due to the low polarity of pristine NPs, 244 their vdW interaction deserves more attention than electrostatic interactions. As a new 245 real space function that is similar to ESP, vdW potential can visually reflect the vdW 246 interaction between a chemical system and the external environment [56]. The 247 advantage of MPI is that it quantifies local polarity induced by the non-uniform 248 distribution of ESP, which is a better descriptor compared to dipole moment that only249 depicts overall polarity.

## 250 **2.4. Weak interaction analysis and visualization**

An independent gradient model based on the Hirshfeld partition (IGMH) of the molecular density method [57] can help to graphically display the weak interactions between two fragments in a complexing system. In order to analyze the nature and region of interaction between NOM and NP fragment, a configuration of a typical NOM–NP dimer was optimized by DFT static calculations in a non-periodic system, and IGMH analysis was performed subsequently by Multifwn 3.8 [53]. All of the isosurface maps were shown and rendered by VMD software [58].

258 The geometry optimization of the dimer was conducted with the 259 CP2K/QUICKSTEP package [59], which was based on a mixed Gaussian and plane 260 waves (GPW) approach [60]. PBE exchange-correlation functional [61], DZVP basis 261 sets [62], and GTH pseudopotentials [63] were applied. The electronic density cutoff 262 was 500 Ry. L-BFGS optimizer [64] was adopted, and convergence criteria for 263 maximum force and displacement were set to be 0.0001 and 0.001 a.u., respectively. Self-consistent continuum solvation (SCCS) implicit solvent model [65] and Grimme's 264 DFT–D3 dispersion correction [52] were employed. 265

### 266 **3. Results and discussion**

### 267 **3.1 Dynamic aggregation processes and structures**

## 268 **3.1.1 Unitary systems**

For unitary systems, the solvent-accessible surface areas (SASAs) as a function of the simulation time of individual NOMs and NPs fragments are shown in Fig. S3. One can observe an overall tendency that all the SASAs are decreasing with simulation time, which indicates the dynamic homo–aggregation processes with the formation of large supramolecular structure (Fig. S4). Notably, the hydrophobic areas of the unaged NPs

274 decreased markedly and remained stable in later simulations, indicating the importance 275 of hydrophobic forces in the self-assembly of the NPs. As NPs molecules come into 276 contact, the water molecules between them are expelled into the bulk due to the weaker 277 interactions between water and hydrophobic surfaces, compared to interactions with 278 other solvents, which is considered an entropy-driven process [66]. Simulation 279 trajectories clearly show that separate NP monomers move toward each other 280 spontaneously and ultimately form a compact cluster. For pristine NPs, it is evident that hydrophobic regions make a dominant contribution to the total SASA, while 281 282 hydrophilic areas make up only a very small proportion (Fig. S3A, C, E), due to the 283 hydrophobic nature of unaged NPs surfaces. Pristine PE homo-aggregate, the least 284 hydrophilic area, is nearly zero (Fig. S3A). In contrast, PVC polymer shows the largest 285 hydrophilic areas, possibly due to the slight polar property induced by Cl atoms in alkyl 286 chains (Fig. S3C), which is consistent with a previous study [44]. In contrast, the 287 hydrophilic areas of aged NPs homogeneous agglomerates (Fig. S3B, D, F) are 288 significantly larger than those of unaged NPs. This phenomenon provides evidence that 289 the oxygen-containing functional groups endow hydrocarbon molecules with partial 290 hydrophilic characteristics. All three types of pristine NPs formed stable aggregate 291 structure of a large aggregate (Fig. S4A, C, E) driven mainly by hydrophobic interaction. 292 Similar to pristine NPs, the supramolecular aggregate of aged NPs homo-293 aggregate is gradually formed with the movement of the whole system. Aged PE 294 ultimately evolved into several small clusters (Fig. S4B) that are less compact compared 295 to the large supramolecular architecture formed by PVC and PS (Fig. S4D and F). 296 Furthermore, the homo-aggregation of aged NPs proceeds through more complicated 297 mechanisms than unaged NPs: hydrophilic (e.g., H-bonding), hydrophobic, and cation 298 bridging (via Ca<sup>2+</sup>) collectively contribute to the homo-aggregation of aged NPs in a 299 water environment. In the unitary system of NOMs, the aggregation of LHA molecules 300 occurs gradually, first by forming close contacts and then by the chelation of functional 301 groups to form a co-aggregate (Fig. S4G), and its aggregation mechanism has been

302 discussed at great length in previous MD studies [3, 67-69].

### 303 3.1.2 Binary systems

304 The root-mean-square deviation (RMSD) curves of NOMs-NPs associations as a 305 function of simulation time are presented in Fig. S5. The curves show that the RMSD 306 values stabilize at approximately 4.5 nm after 40 ns, indicating that the sampling is 307 sufficient. The variations of solvent-accessible surface areas (SASAs) of NOMs-NPs 308 associations over simulation time in binary systems are shown in Fig. 2, which manifest 309 the gradual aggregation processes of NOMs and NPs mixtures for all systems. In the 310 initial configurations, the fragments of NOMs and NPs are arranged in a random 311 manner, and no significant interactions are observed between these fragments. As the system moves, the distances between the NOMs and NPs decrease, driven by the 312 313 gradual increase in intermolecular forces and H-bonding interactions, eventually 314 leading to aggregation. Fig. 3 illustrates the microscopic geometric configurations of 315 the formed NOMs-NPs aggregates. For the NOMs-pristine PE system, all of the PE 316 monomers aggregate to form a small particle with a pillar-like geometry. In this 317 structure, the alkane chains are bundled together and aligned parallel to each other 318 through hydrophobic interactions. Subsequently, NOM fragments or small clusters are 319 adsorbed onto the external surface of the PE polymer through their hydrophobic groups. 320 Such a heterogeneous aggregation mechanism, triggered by the simultaneous 321 adsorption of NOMs and homogeneous aggregation of NPs, results in the formation of 322 a pristine NOMs-NPs PE aggregate (Fig. 3A). In this structure, the outer NOMs expose 323 their hydrophilic functional groups to the external aqueous environment, exhibiting a 324 high binding affinity to metal cations and polar organic substances. In comparison, 325 substantial heterogenous aggregation occurs between NOMs and aged PE fragments 326 due to the increased hydrophilicity of PE. As a result, aged PE and NOMs directly form 327 hetero-aggregating structures through intermolecular interactions and cation bridging 328 (Fig. 3B). For NOMs-pristine PVC binary system, both substances polymerize 329 individually to form separate clusters, and then the two homo-aggregates are bound

330 together primarily through non-polar moieties (Fig. 3C). As shown in Fig. 3D, the case 331 for aged PVC is slightly different: the large cluster of aged PVC and supramolecular 332 aggregate of NOMs have less significant interaction in which a small portion of the 333 deprotonated carboxyl groups of aged PVC and NOMs are bridged by Ca<sup>2+</sup>. For NOMs, which are pristine PS binary systems, the formed agglomerate is structurally similar to 334 335 that of NOMs-pristine PE assemble. That is, PS monomers aggregate by hydrophobic 336 force and intense  $\pi$ - $\pi$  stacking, and then separate NOM clusters with different 337 geometries bind to their outer surface by intermolecular interactions (Fig. 3E). NOMs 338 and aged PS directly form a large heterogeneous aggregate in which fragments of both 339 materials are joined mainly by cation bridging (Fig. 3F), similar to the case for NOMs-340 aged PE aggregate described above. A recent study has demonstrated that as the 341 adsorption time increases, the FTIR peak intensities of PS and PE exhibit an upward 342 trend, while those of PVC show no obvious change, indicating that the binding 343 interaction between PVC and NOMs is different from others [70]. This finding is 344 consistent with the obtained microscopic architectures of NPs-NOMs aggregates. In addition, it has been deduced that if the surface sites of MPs are not rich enough, NOM 345 346 molecules will first occupy the adsorption sites on the surface of pristine MPs and then 347 bind to NOMs already retained on the MP, i.e., the multilayer adsorption mode, but if 348 there are sufficient binding sites on the surface of MPs, HA will tend to be bound as a 349 single layer [44, 71]. In our simulations, the binding of NOMs to NPs shows a structure 350 more resembling multilayer adsorption, possibly due to the similar size of both 351 substances.

To better visualize the microscopic aggregation structures, the solvent water molecules are not shown in the snapshots. In fact, these water molecules are also involved in the formation and stabilization of the supramolecular aggregates of NOMs– NPs associations. Water molecules contribute to the intermolecular attraction between NOMs and NPs fragments, where H–bonds and water bridges formed in both the outer and inner domains of the clusters help to stabilize the hydrophilic functional groups, in 358 good agreement with earlier works [44, 72-74].

359

Fig. 2. Time evolution of solvent accessible surface areas (SASAs) of NOMs–NPs
associations in binary systems.

362

Fig. 3. Snapshots of NOMs–NPs heterogeneous aggregating clusters in binary systems.
Coloring code: N = blue; H = white; O = red; S = yellow; Ca = purple; Cl = orange.
Carbon atoms in NOMs and NPs are depicted in cyan and green, respectively.

366

## 367 **3.2 Interaction energies between NOMs and NPs**

The interaction energies between NOMs and pristine/aged NPs as a function of 368 369 simulation time are provided (Fig. S6) to clearly delineate the fundamental mechanism 370 leading to the spontaneous clustering and assembly of the two substances. For the NOM<sub>s</sub>-pristine PE binary system (Fig. S6A), Coulombic energy is close to zero, 371 372 implying that electrostatic interaction is not favorable for the binding of NOM<sub>S</sub> to PE 373 particle, and in this case, only vdW interaction is responsible for the hetero-aggregation behavior. For NOMs-virgin PVC and PS assemblages, both vdW and electrostatic 374 375 forces have contributions, with vdW interactions significantly stronger than 376 electrostatic interactions, suggesting that vdW interactions are the dominant driving 377 force responsible for aggregation between them and electrostatic interactions play a 378 very limited role; furthermore, electrostatic interaction is slightly more important in the 379 NOMs-PVC than in NOMs-PS system (Fig. S6C and E). In conclusion, vdW 380 interactions are mainly responsible for the binding and aggregation between NOMs and 381 unaged NPs, whereas the electrostatic driving force is only important for PVC. This finding is in good agreement with recent studies [44, 75]. These differences can be 382 383 attributed to the chemical nature of the NP monomers, and our subsequent DFT 384 calculation results will give a direct interpretation of the observed difference.

385 Positive Coulombic interaction energies indicate unfavorable interactions between
386 NOMs and aged NPs (Fig. S6B, D, and F). This is because both substances are

387 negatively charged due to the deprotonation of carboxyl groups, resulting in an overall 388 repulsive force between them. For these systems, vdW interaction energies are slightly 389 negative (Fig. S6B, D, and F), implying that intermolecular forces mildly contribute to 390 the aggregation between NOMs and aged NPs. These results show that the direct interaction between NOMs and aged NPs is not the dominant factor for their 391 aggregation behavior. Instead, the  $Ca^{2+}$  bridging effect attracting and binding to the 392 393 multiple deprotonated carboxylic groups plays a major role in promoting aggregation (discussed below). 394

## **395 3.3. Local coordination by cation bridging**

The bridging effect of cations such as  $Ca^{2+}$  has been recognized as a fundamental 396 397 mechanism for the stabilization and aggregation of NOM fragments that are abundant in deprotonated carboxyl groups in near-neutral pH environments [5, 29, 76, 77]. In 398 399 comparison, the bridging effect between carboxyl groups of NOMs by Na<sup>+</sup> is relatively weak, which means the aggregation of NOMs is not effectively promoted by Na<sup>+</sup>. 400 Consequently,  $Ca^{2+}$  cations in a water environment mainly contribute to the aggregation 401 behavior of NOMs and aged NPs. To quantitatively characterize the local coordination 402 structures of Ca<sup>2+</sup> complexation with deprotonated carboxyl groups (COO<sup>-</sup>) and to 403 distinguish Ca-bridged homogeneous and heterogeneous aggregation, radial 404 405 distribution functions (RDFs) were analyzed and presented.

Fig. S7 shows the RDFs curves of aqueous Ca<sup>2+</sup> with respect to carboxyl O atoms 406 of NOMs and aged NPs, respectively, in binary systems. One can find that the scenarios 407 of Ca<sup>2+</sup> coordination with carboxyl O of NOMs are very similar in all of the three 408 409 systems (Fig. S7A, C, and E). In brief, the RDF curves show a sharp and symmetrical peak centered at ~2.60 Å, which corresponds to the average Ca–O <sup>carboxyl</sup> bond length, 410 and the accumulating coordination number (CN) of the peak is around 2.1, which 411 indicates that, on average, each  $Ca^{2+}$  in solution binds firmly with 2.1 O atoms from 412 NOMs' carboxyl groups. Moreover, in the three binary systems, Ca<sup>2+</sup> is respectively 413 bound to 0.92, 0.50, and 0.38 carboxyl O of aged PE, PS, and PVC with an average 414

distance of 2.60 Å (Fig. S7B, D, and F). These data show that when NOMs and aged 415 NPs coexist, homogeneous aggregation of both substances still occurs by  $Ca^{2+}$  bridging. 416 Fig. 4 shows the RDF and CN profiles of the C carboxyl of NOMs surrounded by the 417 C carboxyl of aged NPs, which helps to better reveal the heterogeneous aggregation 418 between them formed by Ca2+ bridging, and also illustrates representative 419 420 configurations corresponding to the peak positions of the RDF to visually show local 421 polymerizing structures of NOMs-NPs heterogeneous aggregation. A distinct peak or multiple peaks between 3 and 6 Å can be identified in the carboxyl-carboxyl RDF 422 curves (Fig. 4A–C), and the accumulated CN values (at 6.0 Å) reveal that, on average, 423 424 each carboxyl group in NOMs is associated with 0.27, 0.16, and 0.10 carboxyl groups in aged PE, PS, and PVC, forming multiple carboxyl chelates bridged by  $Ca^{2+}$  (Fig. 425 4D–F) and finally evolving into a heterogeneous aggregate. 426

To summarize,  $Ca^{2+}$  plays a crucial role in the aggregation (homogeneous and heterogeneous) and stabilization of NOMs and aged NPs because the negatively charged carboxyl groups are partially neutralized by  $Ca^{2+}$  cations in solution, which reduces the intermolecular and intramolecular charge repulsion and limits conformational expansion, thereby initializing the formation of compact aggregates.

432

Fig. 4. Radial distribution function (RDF) and coordination number (CN) curves of the carboxyl C of NOMs with respect to the carboxyl C of aged NPs (A–C) and representative hetero–polymerization structures bridged by  $Ca^{2+}$  (D–F). N = blue; H = white; O = red; Ca = purple; Cl = orange. C atoms in NOMs and NPs are shown in cyan and green, respectively.

438

## 439 **3.4.** Quantitative analysis of molecular surface by DFT calculations

440 **3.4.1. Electrostatic potential** 

The isosurface-colored maps of electrostatic potential (ESP) (Fig. 5) are used to characterize the overall ESP distributions of the six monomeric species of NPs. For virgin PE, it can be found that due to its non-polar geometry, its vdW surface distributes

ESP of nearly zero evenly (Fig. 5A), resulting in the smallest magnitude of ESP among all the NPs. Virgin PVC exhibits a significantly higher ESP scene, with the regions around the Cl atom showing minimum negative ESP and the domains surrounding the H atom presenting maximum positive values (Fig. 5B). Overall, virgin PS has a slightly higher electronegativity than PE and the lowest ESPs, characterized by an annular negative isosurface, are found in the regions between the benzene rings, while other regions exhibit a significantly small magnitude of ESP (Fig. 5C).

451 Compared to the pristine NPs, the surface ESPs of the aged NPs are distributed in a 452 significantly lower region due to the introduction of non-uniformly distributed oxygen-453 containing moieties, which are found to increase the non-uniform distribution of the 454 surface ESPs. This, together with the increased H-bonding interactions, promotes the increase in adsorption and aggregation with NOMs. It can be clearly seen that the most 455 negative ESP values are caused by the deprotonated carboxylic groups in all three aged 456 NPs (Fig. 5D-F), indicating that these sites are prone to forming pronounced 457 electrostatic interactions with cations. 458

459

Fig. 5. Isosurface maps of the electrostatic potential (ESP) for (A) pristine PE, (B) pristine PVC, (C) pristine PS, (D) aged PE, (E) aged PVC, and (F) aged PS (isosurface = 0.001 a.u.). Red and blue colors correspond to the positive and negative parts of ESP, respectively. Isosurface maps of vdW potential for (G) pristine PE, (H) pristine PVC, and (I) pristine PS (isovalue =  $\pm 0.75$  kcal/mol). Green and blue colors correspond to positive and negative values, respectively.

466

## 467 **3.4.2. vdW potential**

Due to the low polarity of pristine NPs, their vdW interaction should be of greater importance in comparison to their electrostatic interaction according to chemical intuition. Consequently, the isosurface maps of the vdW potentials of the three pristine NPs are depicted. (Fig. 5G–I). This study primarily focused on the blue regions, where the effect of dispersive attraction outweighs that of exchange–repulsion, and generally, the interacting atoms bind to such region. For unaged PE, the negative part of its vdW

474 potential isosurface (blue region) is symmetrically distributed on both sides around C 475 atoms (Fig. 5G). Due to the negative ESP and the polarity generated by the Cl atoms in 476 virgin PVC, the blue region of its vdW potential isosurface is mainly concentrated 477 around C (Fig. 5H). For pristine PS, the blue region is predominantly near the ring 478 center (Fig. 51). The strong dispersive attraction in this region suggests that  $\pi - \pi$  stacking 479 interaction plays a pivotal role in its aggregation with NOMs. In fact, strong  $\pi - \pi$ 480 interactions between PS and NOMs are clearly identified in the MD trajectories. An 481 earlier study also suggested the  $\pi$ - $\pi$  conjugation for PS interaction with NOMs [23].

### **4**82 **3**

### **3.4.3. Molecular polarity index (MPI)**

MPI is used to quantify the polarity of molecules with non-uniformly distributed 483 484 ESP. MPI is calculated by averaging the absolute values of ESPs on the surface of a 485 molecule. A higher MPI indicates a stronger polarity and a greater binding affinity 486 through electrostatic interactions. The calculated MPI values of the six NP monomers 487 are provided in Table S3, and for comparison, the MPI of water molecules is also 488 computed at the same theoretical level. The MPIs for unaged PE, PS, and PVC are 3.1, 489 8.5, and 22.2 kcal/mol, respectively. Therefore, PE exhibits the lowest polarity, and PS 490 also presents a very low polarity. As a result, it is mainly the vdW interactions, rather 491 than electrostatic interactions, that drive the aggregation of virgin PE and PS. In contrast, 492 the aged NPs are much more polar than their pristine forms, as the derived MPIs are 493 43.2 kcal/mol for aged PE, 51.6 kcal/mol for aged PS, and 42.2 kcal/mol for PVC. The 494 MPI of the H<sub>2</sub>O molecule is found to be 22.1 kcal/mol, which is similar to that of 495 pristine PVC (22.2 kcal/mol) and much higher than that of pristine PE (3.1 kcal/mol) 496 and PS (8.5 kcal/mol). The MPI of H<sub>2</sub>O is significantly lower than that of aged NPs (at 497 least 42.2 kcal/mol), and therefore the polarity of the water molecule is much smaller than that of aged NPs, confirming that the strength of electrostatic interactions between 498 499 aged NPs and NOMs is rather weak.

## 500 **3.5. Visualization analysis of weak interactions**

501 Based on DFT calculations and wavefunction analysis, IGMH is a highly versatile 502 and effective analysis method for identifying and visualizing areas of significant 503 interaction in a wide range of chemical systems. An outstanding advantage of IGMH is 504 the ability to define fragments, so as to exclusively analyze the intermolecular 505 interactions between fragments. Taking the NOM-aged PS system as an example, the 506 DFT geometry optimization method was used to obtain the structure of NOM-aged PS 507 dimer with the initial configuration (1 aged PS molecule + 1 NOM fragment) from the 508 trajectory of MD simulations. According to the acquired structure, IGMH analysis of 509 the two fragments in such a dimer was carried out.

510 It can be seen that the isosurfaces between aged PS and NOM are largely green (Fig. 6), which implies that slightly low electron density is in the region related to 511 512 intermolecular interaction, and the molecular binding between the two fragments is 513 predominantly induced by dispersion attraction, since electrostatic interaction normally is accompanied by relatively higher electron density. Although the dispersion 514 515 interaction between atom pairs is typically assumed to be weak, the isosurfaces between 516 the fragments are notably wide, almost covering the whole interaction region (Fig. 6). 517 This observation could suggest that the vdW interaction between NOM and aged PS 518 fragments is not weak and play a significant role in the binding. Moreover, the 519 isosurfaces located between the rings of NOM and aged PS are stacked parallelly (Fig. 520 6), which signifies a significant  $\pi$ - $\pi$  interaction. The  $\pi$ - $\pi$  stacking effect is expected to 521 increase the binding strength. However, due to the big structural difference, there is no 522 comparable binding mode or interaction mechanism for PE and PVC. Hence, it can be 523 deduced that for other kinds of NPs that possess a benzene ring,  $\pi$ - $\pi$  interaction is likely 524 to noticeably intensify their adhesion and aggregation with NOMs.

525

526 Fig. 6. Sign( $\lambda_2$ ) $\rho$  colored IGMH isosurface maps (of NOM-aged PS dimer. The NOM 527 monomer and PS monomer are defined as the two fragments in IGMH analysis.  $\delta g^{inter}$ 528 = 0.003 a.u.

## 529 **4.** Conclusion

530 This study investigated the dynamic aggregation and binding mechanisms between NOMs and NPs at the molecular scale. Molecular modeling results showed 531 532 that the virgin NPs tend to form a homogeneous assembly, which is mainly driven by 533 hydrophobic force and vdW interaction. In contrast, due to the presence of negatively 534 charged carboxyl groups, the homogeneous aggregation of aged NPs is mainly caused 535 by the cation bridging effect, which partially neutralizes the negatively charged 536 carboxylic groups. Pristine PE and PS can easily assemble into a homogenous aggregate, 537 and then NOM molecules tend to be adsorbed on the outer surfaces of such NP particles, 538 then the subsequent NOMs can either bind to the NP surface sites or adhere to the NOM 539 fragments that have already been attached to the NPs. In comparison, NOM fragments 540 are not obviously adsorbed on the homogenous aggregate of virgin PVC due to its 541 relatively higher polarity and the presence of chlorine; instead, NOM molecules tend to 542 self-aggregate and then come into contact with the pristine PVC particle to form an 543 interface. In the case of aged NPs, their interactions with NOMs are governed by a 544 variety of mechanisms, such as intermolecular hydrogen bonding involving hydrophilic 545 functional groups, hydrophobic interactions including  $\pi$ - $\pi$  stacking, and the cation 546 bridging effect.

547

## 548 **CRediT authorship contribution statement**

C.Z.: conceptualization, writing–original draft, writing–review & editing, investigation.
Z.Y.Z.: writing–original draft, writing–review & editing. M.N.X.: data curation,
software. H.Z.M.: data analysis. J.H.Q.: writing–review & editing, supervision,
resources. H.Z.J.: conceptualization, writing–review & editing, supervision, resources.

## 553 Declaration of Competing Interest

554 The authors declare that they have no known competing financial interests or personal 555 relationships that could have appeared to influence the work reported in this paper.

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- 561

## 562 **References**

- 563 [1] Kellerman, A. M.; Kothawala, D. N.; Dittmar, T.; Tranvik, L. J., Persistence of
  564 dissolved organic matter in lakes related to its molecular characteristics. *Nat.*565 *Geosci.* 2015, *8*, (6), 454-457.
- 566 [2] Lehmann, J.; Kleber, M., The contentious nature of soil organic matter. *Nature*567 **2015**, *528*, (7580), 60-68.
- [3] Zhou, Z.; Zhang, C.; Xi, M.; Ma, H.; Jia, H., Multi-scale modeling of natural
  organic matter-heavy metal cations interactions: Aggregation and stabilization
  mechanisms. *Water Res.* 2023, 238, 120007.
- [4] Lan, T.; Wu, P.; Yin, X.; Zhao, Y.; Liao, J.; Wang, D.; et al., Rigidity and Flexibility:
  Unraveling the Role of Fulvic Acid in Uranyl Sorption on Graphene Oxide Using
  Molecular Dynamics Simulations. *Environ. Sci. Technol.* 2023, *57*, (28), 1033910347.
- 575 [5] Petrov, D.; Tunega, D.; Gerzabek, M. H.; Oostenbrink, C., Molecular Dynamics
  576 Simulations of the Standard Leonardite Humic Acid: Microscopic Analysis of the
  577 Structure and Dynamics. *Environ. Sci. Technol.* 2017, *51*, (10), 5414-5424.
- 578 [6] Ali, I.; Tan, X.; Li, J.; Peng, C.; Naz, I.; Duan, Z.; et al., Interaction of microplastics
  579 and nanoplastics with natural organic matter (NOM) and the impact of NOM on
  580 the sorption behavior of anthropogenic contaminants A critical review. *Journal*581 of Cleaner Production 2022, 376, 134314.
- Jambeck, J. R.; Geyer, R.; Wilcox, C.; Siegler, T. R.; Perryman, M.; Andrady, A.;
  et al., Plastic waste inputs from land into the ocean. *Science* 2015, *347*, (6223),
  768-771.
- 585 [8] Gewert, B.; Plassmann, M. M.; MacLeod, M., Pathways for degradation of plastic
  586 polymers floating in the marine environment. *Environ Sci Process Impacts* 2015,
  587 17, (9), 1513-21.
- 588 [9] Moharir, R. V.; Kumar, S., Challenges associated with plastic waste disposal and
  allied microbial routes for its effective degradation: A comprehensive review. *Journal of Cleaner Production* 2019, 208, 65-76.
- [10] Hu, D.; Shen, M.; Zhang, Y.; Zeng, G., Micro(nano)plastics: An un-ignorable
  carbon source? *Sci. Total Environ.* 2019, 657, 108-110.
- [11] Liu, G.; Zhu, Z.; Yang, Y.; Sun, Y.; Yu, F.; Ma, J., Sorption behavior and
   mechanism of hydrophilic organic chemicals to virgin and aged microplastics in

freshwater and seawater. *Environ. Pollut.* **2019**, *246*, 26-33.

- 596 [12] Wang, Q.; Zhang, Y.; Wangjin, X.; Wang, Y.; Meng, G.; Chen, Y., The adsorption
  597 behavior of metals in aqueous solution by microplastics effected by UV radiation.
  598 *J. Environ. Sci. (China)* 2020, *87*, 272-280.
- [13]Shen, M.; Zhu, Y.; Zhang, Y.; Zeng, G.; Wen, X.; Yi, H.; et al., Micro(nano)plastics:
  Unignorable vectors for organisms. *Mar. Pollut. Bull.* 2019, *139*, 328-331.
- [14] Pradel, A.; Catrouillet, C.; Gigault, J., The environmental fate of nanoplastics:
  What we know and what we need to know about aggregation. *NanoImpact* 2023,
  29, 100453.
- [15] Alimi, O. S.; Farner Budarz, J.; Hernandez, L. M.; Tufenkji, N., Microplastics and
  Nanoplastics in Aquatic Environments: Aggregation, Deposition, and Enhanced
  Contaminant Transport. *Environ. Sci. Technol.* 2018, *52*, (4), 1704-1724.
- 607 [16] Gigault, J.; El Hadri, H.; Nguyen, B.; Grassl, B.; Rowenczyk, L.; Tufenkji, N.; et
  608 al., Nanoplastics are neither microplastics nor engineered nanoparticles. *Nature*609 *Nanotechnology* 2021, *16*, (5), 501-507.
- [17] Lambert, S.; Wagner, M., Characterisation of nanoplastics during the degradation
  of polystyrene. *Chemosphere* 2016, *145*, 265-268.
- [18] Liu, Y.; Wang, Y.; Ling, X.; Yan, Z.; Wu, D.; Liu, J.; et al., Effects of Nanoplastics
  and Butyl Methoxydibenzoylmethane on Early Zebrafish Embryos Identified by
  Single-Cell RNA Sequencing. *Environ. Sci. Technol.* 2021, 55, (3), 1885-1896.
- [19] Wang, Z.; Xing, X.; Xue, M.; Bai, S.; Li, P.; Li, C.; et al., Insights into
  heteroaggregation of polystyrene nanoplastics with hematite nanoparticles and
  configuration-dependent adsorption for PFOA and PFOS. *Colloids Surf., A* 2022,
  618 649, 129467.
- [20] Zhang, Y.; Luo, Y.; Yu, X.; Huang, D.; Guo, X.; Zhu, L., Aging significantly
  increases the interaction between polystyrene nanoplastic and minerals. *Water Res.* **2022**, *219*, 118544.
- [21] Mao, Y.; Li, H.; Huangfu, X.; Liu, Y.; He, Q., Nanoplastics display strong stability
  in aqueous environments: Insights from aggregation behaviour and theoretical
  calculations. *Environ. Pollut.* 2020, 258, 113760.
- [22] Singh, N.; Tiwari, E.; Khandelwal, N.; Darbha, G. K., Understanding the stability
  of nanoplastics in aqueous environments: effect of ionic strength, temperature,
  dissolved organic matter, clay, and heavy metals. *Environmental Science: Nano*2019, 6, (10), 2968-2976.
- [23] Ding, L.; Luo, Y.; Yu, X.; Ouyang, Z.; Liu, P.; Guo, X., Insight into interactions of
  polystyrene microplastics with different types and compositions of dissolved
  organic matter. *Sci. Total Environ.* 2022, *824*, 153883.
- [24] Yin, Y.; Shen, M.; Tan, Z.; Yu, S.; Liu, J.; Jiang, G., Particle coating-dependent
  interaction of molecular weight fractionated natural organic matter: impacts on the
  aggregation of silver nanoparticles. *Environ. Sci. Technol.* 2015, 49, (11), 6581-9.
- 635 [25] Oriekhova, O.; Stoll, S., Heteroaggregation of nanoplastic particles in the presence
   636 of inorganic colloids and natural organic matter. *Environmental Science: Nano*

- **2018**, *5*, (3), 792-799.
- 638 [26] Greathouse, J.; Johnson, K.; Greenwell, H., Interaction of Natural Organic Matter
  639 with Layered Minerals: Recent Developments in Computational Methods at the
  640 Nanoscale. *Minerals* 2014, 4, (2), 519-540.
- [27] Devarajan, D.; Liang, L.; Gu, B.; Brooks, S. C.; Parks, J. M.; Smith, J. C.,
  Molecular Dynamics Simulation of the Structures, Dynamics, and Aggregation of
  Dissolved Organic Matter. *Environ. Sci. Technol.* 2020, *54*, (21), 13527-13537.
- [28] Kalinichev, A. G.; Iskrenova-Tchoukova, E.; Ahn, W. Y.; Clark, M. M.;
  Kirkpatrick, R. J., Effects of Ca<sup>2+</sup> on supramolecular aggregation of natural
  organic matter in aqueous solutions: A comparison of molecular modeling
  approaches. *Geoderma* 2011, *169*, 27-32.
- [29] Zhang, Y.; Liu, X.; Zhang, C.; Lu, X., A combined first principles and classical
  molecular dynamics study of clay-soil organic matters (SOMs) interactions. *Geochim. Cosmochim. Acta* 2020, 291, 110-125.
- [30] Liang, J.; Zhen, P.; Gan, P.; Li, Y.; Tong, M.; Liu, W., DFT Calculation of
  Nonperiodic Small Molecular Systems to Predict the Reaction Mechanism of
  Advanced Oxidation Processes: Challenges and Perspectives. ACS ES&T *Engineering* 2023.
- [31] Li, F.; Borthwick, A. G. L.; Liu, W., Environmental theoretical calculation for non periodic systems. *Trends in Chemistry* 2023, 5, (6), 410-414.
- [32] He, M.; Liu, X.; Lu, X.; Zhang, Y.; Wang, R., Structure, Stability, and Acidity of
  the Uranyl Arsenate Dimer in Aqueous Solution. *Inorg. Chem.* 2023, *62*, (22),
  8729-8738.
- [33] He, M.; Liu, X.; Cheng, J.; Lu, X.; Zhang, C.; Wang, R., Uranyl Arsenate
  Complexes in Aqueous Solution: Insights from First-Principles Molecular
  Dynamics Simulations. *Inorg. Chem.* 2018, 57, (10), 5801-5809.
- [34] Xue, Q.; Jiao, Z.; Pan, W.; Liu, X.; Fu, J.; Zhang, A., Multiscale computational
  simulation of pollutant behavior at water interfaces. *Water Res.* 2024, 250, 121043.
- [35] Chen, Y.; Li, H.; Yin, Y.; Shan, S.; Huang, T.; Tang, H., Effect of microplastics on
  the adherence of coexisting background organic contaminants to natural organic
  matter in water. *Sci. Total Environ.* 2023, *905*, 167175.
- [36] Zhu, K.; Jia, H.; Zhao, S.; Xia, T.; Guo, X.; Wang, T.; et al., Formation of
  Environmentally Persistent Free Radicals on Microplastics under Light Irradiation. *Environ. Sci. Technol.* 2019, *53*, (14), 8177-8186.
- [37] Luo, Y.; Zhang, Y.; Xu, Y.; Guo, X.; Zhu, L., Distribution characteristics and
  mechanism of microplastics mediated by soil physicochemical properties. *Sci. Total Environ.* 2020, 726, 138389.
- [38] Zhu, K.; Sun, Y.; Jiang, W.; Zhang, C.; Dai, Y.; Liu, Z.; et al., Inorganic anions
  influenced the photoaging kinetics and mechanism of polystyrene microplastic
  under the simulated sunlight: Role of reactive radical species. *Water Res.* 2022,
  216, 118294.
- 678 [39] Escalona, Y.; Petrov, D.; Oostenbrink, C., Vienna soil organic matter modeler 2

679	(VSOMM2). J. Mol. Graphics Modell. 2021, 103, 107817.
680	[40] Sündermann, A.; Solc, R.; Tunega, D.; Haberhauer, G.; Gerzabek, M. H.;
681	Oostenbrink, C., Vienna Soil-Organic-Matter Modeler-Generating condensed-
682	phase models of humic substances. J. Mol. Graphics Modell. 2015, 62, 253-261.
683	[41] Chen, Y.; Tang, H.; Cheng, Y.; Huang, T.; Xing, B., Interaction between
684	microplastics and humic acid and its effect on their properties as revealed by
685	molecular dynamics simulations. J. Hazard. Mater. 2023, 455, 131636.
686	[42] Van Der Spoel, D.; Lindahl, E.; Hess, B.; Groenhof, G.; Mark, A. E.; Berendsen,
687	H. J., GROMACS: fast, flexible, and free. J. Comput. Chem. 2005, 26, (16), 1701-
688	18.
689	[43] Duan, Y.; Wu, C.; Chowdhury, S.; Lee, M. C.; Xiong, G.; Zhang, W.; et al., A point-
690	charge force field for molecular mechanics simulations of proteins based on
691	condensed-phase quantum mechanical calculations. J. Comput. Chem. 2003, 24,
692	(16), 1999-2012.
693	[44] Essmann, U.; Perera, L.; Berkowitz, M. L.; Darden, T.; Lee, H.; Pedersen, L. G.,
694	A smooth particle mesh Ewald method. J. Chem. Phys. 1995, 103, (19), 8577-8593.
695	[45] Berendsen, H. J. C.; Grigera, J. R.; Straatsma, T. P., The missing term in effective
696	pair potentials. J. Phys. Chem. C 1987, 91, (24), 6269-6271.
697	[46] Hess, B.; Bekker, H.; Berendsen, H. J. C.; Fraaije, J. G. E. M., LINCS: A linear
698	constraint solver for molecular simulations. <b>1997</b> , <i>18</i> , (12), 1463-1472.
699	[47] Berendsen; Postma, J. P. M.; van Gunsteren, W. F.; DiNola, A.; Haak, J. R.,
700	Molecular dynamics with coupling to an external bath. J. Chem. Phys. 1984, 81,
701	(8), 3684-3690.
702	[48] Bussi, G.; Donadio, D.; Parrinello, M., Canonical sampling through velocity
703	rescaling. J. Chem. Phys. 2007, 126, (1), 014101.
704	[49] Weigend, F.; Ahlrichs, R., Balanced basis sets of split valence, triple zeta valence
705	and quadruple zeta valence quality for H to Rn: Design and assessment of accuracy.
706	Phys. Chem. Chem. Phys. 2005, 7, (18), 3297-305.
707	[50] Becke, A. D., Density-functional thermochemistry. III. The role of exact exchange.
708	J. Chem. Phys. 1993, 98, (7), 5648-5652.
709	[51] Lee, C.; Yang, W.; Parr, R. G., Development of the Colle-Salvetti correlation-
710	energy formula into a functional of the electron density. Phys Rev B Condens
711	Matter 1988, 37, (2), 785-789.
712	[52] Grimme, S.; Antony, J.; Ehrlich, S.; Krieg, H., A consistent and accurate ab initio
713	parametrization of density functional dispersion correction (DFT-D) for the 94
714	elements H-Pu. J. Chem. Phys. 2010, 132, (15), 154104.
715	[53] Lu, T.; Chen, F., Multiwfn: a multifunctional wavefunction analyzer. J. Comput.
716	<i>Chem.</i> <b>2012,</b> <i>33</i> , (5), 580-92.
717	[54] Schauperl, M.; Nerenberg, P. S.; Jang, H.; Wang, LP.; Bayly, C. I.; Mobley, D.
718	L.; et al., Non-bonded force field model with advanced restrained electrostatic
719	potential charges (RESP2). Communications Chemistry 2020, 3, (1), 44.
720	[55] Bader, R. F. W.; Carroll, M. T.; Cheeseman, J. R.; Chang, C., Properties of atoms

721	in molecules: atomic volumes. J. Am. Chem. Soc. 1987, 109, (26), 7968-7979.
722	[56] Lu, T.; Chen, Q., van der Waals potential: an important complement to molecular
723	electrostatic potential in studying intermolecular interactions. J. Mol. Model. 2020,
724	26, (11), 315.
725	[57] Lu, T.; Chen, Q., Independent gradient model based on Hirshfeld partition: A new
726	method for visual study of interactions in chemical systems. 2022, 43, (8), 539-
727	555.
728	[58] Humphrey, W.; Dalke, A.; Schulten, K., VMD: Visual molecular dynamics. J. Mol.
729	<i>Graphics</i> <b>1996,</b> <i>14</i> , (1), 33-38.
730	[59] Kuehne, T. D.; Iannuzzi, M.; Del Ben, M.; Rybkin, V. V.; Seewald, P.; Stein, F.; et
731	al., CP2K: An electronic structure and molecular dynamics software package -
732	Quickstep: Efficient and accurate electronic structure calculations. J. Chem. Phys.
733	<b>2020,</b> <i>152</i> , (19).
734	[60] Lippert, B. G.; Hutter, J.; Parrinello, M., A hybrid Gaussian and plane wave density
735	functional scheme. Mol. Phys. 1997, 92, (92), 477-487.
736	[61] Perdew, J. P.; Burke, K.; Ernzerhof, M., Generalized gradient approximation made
737	simple. Phys. Rev. Lett. 1996, 77, (18), 3865-3868.
738	[62] VandeVondele, J.; Hutter, J., Gaussian basis sets for accurate calculations on
739	molecular systems in gas and condensed phases. J. Chem. Phys. 2007, 127, (11),
740	114105.
741	[63] Goedecker, S.; Teter, M.; Hutter, J., Separable dual-space Gaussian
742	pseudopotentials. Phys. Rev. B: Condens. Matter 1996, 54, (3), 17031710.
743	[64] Byrd, R. H.; Lu, P.; Nocedal, J.; Zhu, C., A Limited Memory Algorithm for Bound
744	Constrained Optimization. 1995, 16, (5), 1190-1208.
745	[65] Andreussi, O.; Dabo, I.; Marzari, N., Revised self-consistent continuum solvation
746	in electronic-structure calculations. J. Chem. Phys. 2012, 136, (6).
747	[66] Sun, Q., The Hydrophobic Effects: Our Current Understanding. <i>Molecules</i> 2022,
748	27, (20).
749	[67] Aquino, A. J. A.; Tunega, D.; Schaumann, G. E.; Haberhauer, G.; Gerzabek, M.
750	H.; Lischka, H., The functionality of cation bridges for binding polar groups in
751	soil aggregates. Int. J. Quantum Chem 2010, 111, (7-8), 1531-1542.
752	[68] Galicia-Andrés, E.; Oostenbrink, C.; Gerzabek, M. H.; Tunega, D., On the
753	Adsorption Mechanism of Humic Substances on Kaolinite and Their Microscopic
754	Structure. <i>Minerals</i> <b>2021</b> , <i>11</i> , (10).
755	[69] Escalona, Y.; Petrov, D.; Galicia-Andrés, E.; Oostenbrink, C., Exploring the
756	Macroscopic Properties of Humic Substances Using Modeling and Molecular
757	Simulations. <i>Agronomy</i> <b>2023</b> , <i>13</i> , (4).
758	[70] Wang, X.; Wang, X.; Zhu, W.; Ding, L.; Liang, X.; Wu, R.; et al., Insight into
759	interactions between microplastics and fulvic acid: Mechanisms affected by
760	microplastics type. Sci. Total Environ. 2024, 913, 169427.
761	[71] Abdurahman, A.; Cui, K.; Wu, J.; Li, S.; Gao, R.; Dai, J.; et al., Adsorption of
762	dissolved organic matter (DOM) on polystyrene microplastics in aquatic

763	environments: Kinetic, isotherm and site energy distribution analysis. Ecotoxicol.
764	Environ. Saf. 2020, 198, 110658.
765	[72] Aquino, A. J. A.; Tunega, D.; Pašalić, H.; Schaumann, G. E.; Haberhauer, G.;
766	Gerzabek, M. H.; et al., Molecular Dynamics Simulations of Water Molecule-
767	Bridges in Polar Domains of Humic Acids. Environ. Sci. Technol. 2011, 45, (19),
768	8411-8419.
769	[73] Kučerík, J.; Ondruch, P.; Kunhi Mouvenchery, Y.; Schaumann, G. E., Formation
770	of Water Molecule Bridges Governs Water Sorption Mechanisms in Soil Organic
771	Matter. Langmuir 2018, 34, (40), 12174-12182.
772	[74] Boyle, M.; Frankenberger Jr., W. T.; Stolzy, L. H., The Influence of Organic Matter
773	on Soil Aggregation and Water Infiltration. 1989, 2, (4), 290-299.
774	[75] Liu, W.; Tang, H.; Yang, B.; Li, C.; Chen, Y.; Huang, T., Molecular level insight
775	into the different interaction intensity between microplastics and aromatic
776	hydrocarbon in pure water and seawater. Sci. Total Environ. 2023, 862, 160786.
777	[76] Iskrenova-Tchoukova, E.; Kalinichev, A. G.; Kirkpatrick, R. J., Metal Cation
778	Complexation with Natural Organic Matter in Aqueous Solutions: Molecular
779	Dynamics Simulations and Potentials of Mean Force. Langmuir 2010, 26, (20),
780	15909-15919.
781	[77] Kalinichev, A. G.; Kirkpatrick, R. J., Molecular dynamics simulation of cationic
782	complexation with natural organic matter. <b>2007</b> , <i>58</i> , (4), 909-917.

A: NOMs-unaged NPs

B: NOMs-aged NPs









E : NOMs-unaged PS



B : NOMs-aged PE



D : NOMs-aged PVC



F : NOMs-aged PS







F : NOMs-aged PVC association







## Highlights:

· NOMs-NPs interactions were revealed by multi-scale computational modeling

• Advanced NOMs models created by VSOMM2 were used in simulations.

• Dynamic aggregation mechanisms were elucidated and quantified.

• Pristine and aged NPs show dissimilar aggregation mechanisms.

• PE, PS, and PVC aggregation with NOMs occur via different processes.

<text>