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*In situ* imaging of microplastics in living organisms based on mass spectrometry technology

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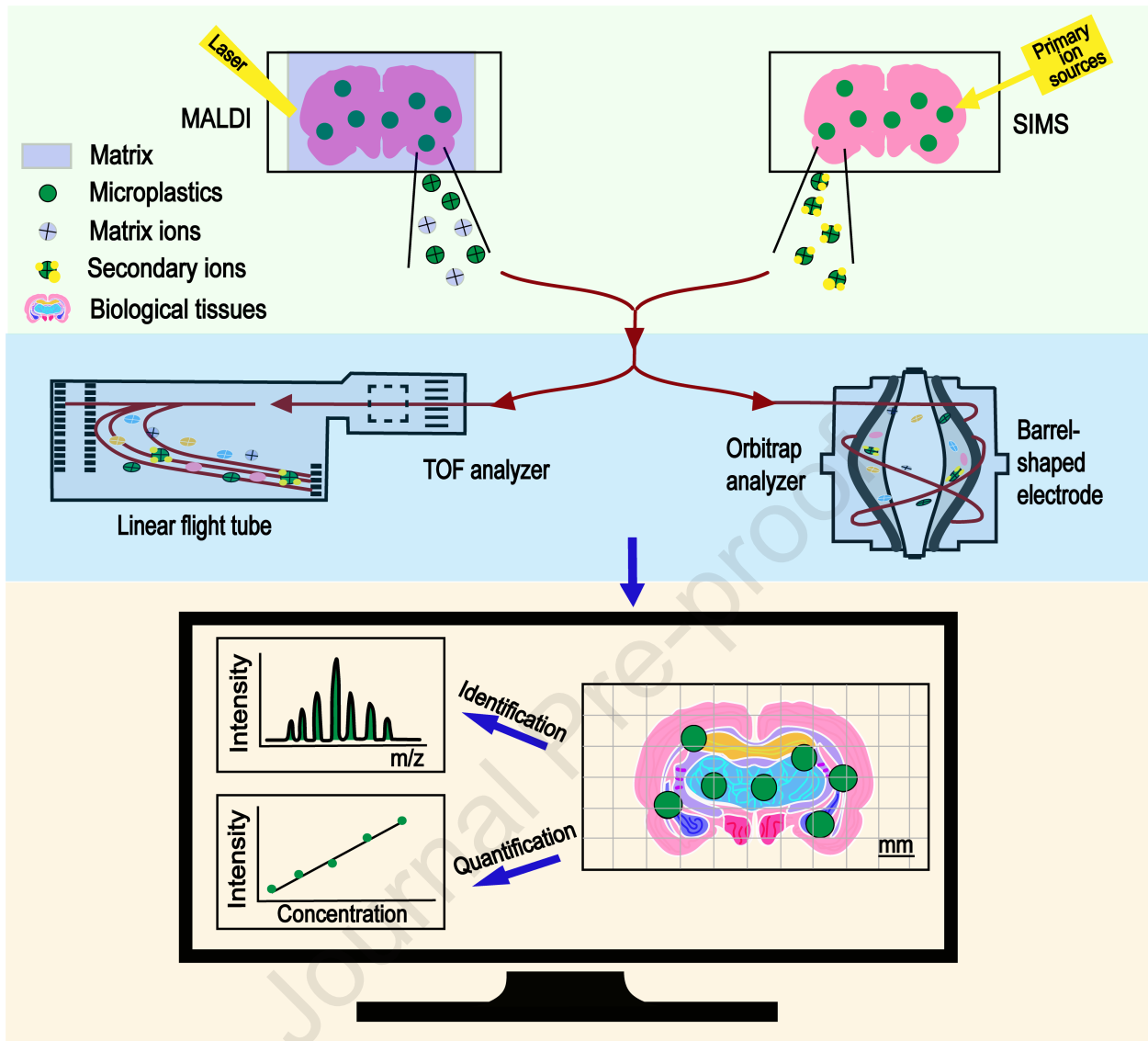
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1 ***In situ* imaging of microplastics in living organisms based on mass spectrometry**  
2 **technology**

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17

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20 **Abstract:**

21 Plastic pollution is widely present in terrestrial and aquatic ecosystems, and  
22 microplastics (MPs) can be detected in organisms. *In situ* detection methods for MPs  
23 in organisms have attracted widespread attention. Traditional imaging characterization  
24 methods of MPs, including stereo microscopes and fluorescence microscopy, are  
25 typically used to image artificially added microsphere standards under laboratory  
26 conditions. However, they cannot specifically identify MPs in biological samples. Thus,  
27 there is a need for a detection technique that can provide spatial distribution information  
28 of MPs in biological samples, as well as measure their quality and quantity. In this  
29 perspective, to obtain high-resolution images with chemical composition analysis, we  
30 compared ion sources for ionizing plastic macromolecules and mass analyzers for  
31 analyzing macromolecules. Matrix-assisted laser desorption/ionization (MALDI) is  
32 suitable for imaging characterization, while time-of-flight (TOF) and Orbitrap mass  
33 spectrometry are suitable for polymer mass spectrometry analysis. Furthermore, we  
34 propose a technique that combines MALDI with TOF or Orbitrap, which holds promise  
35 for the *in situ* imaging of MPs in biological samples.

36

37 **Keywords:** Microplastics; Organisms; *in situ* imaging; Mass spectrometry imaging

38

## 39 1. Introduction

40 Microplastics (MPs) are widely present in terrestrial and aquatic environments [1]  
41 in the form of fragments, fibers, and films with a diameter less than 5 mm [2, 3]. Studies  
42 have shown that MPs also exist in organisms, such as the guts and livers of fish [4, 5],  
43 the roots and stems of plants [6-8], and human blood [9, 10]. Besides, submicroplastics  
44 (100 nm to 1  $\mu$ m) and nanoplastics (< 100 nm) have been shown to penetrate the blood-  
45 brain barrier of fishes [11, 12], the placental barrier of humans [13, 14], the skin barrier of  
46 humans and mice [15, 16], and undergo internalization by bovine oviductal epithelial cells  
47 and human colon fibroblasts [17]. Therefore, The MPs in organisms need to be  
48 characterized and detected.

49 The detection methods for MPs are divided into qualitative methods and  
50 quantitative methods [18]. Qualitative detection aims to confirm the existence of MPs  
51 and characterize the type, morphology, and size of MPs. The coupling of microscopes  
52 with vibrational spectroscopy techniques to identify MPs is the most common approach  
53 [19]. Quantitative detection of MPs in terms of quantity can also be performed using  
54 microscopic imaging. Furthermore, the distribution characteristics of fluorescently  
55 stained or radiolabeled MPs in organisms can be characterized by the intensity of  
56 fluorescence or radioactivity. Mass spectrometry (MS) detection can provide mass-  
57 related information about MPs, but it cannot simultaneously obtain their morphology,  
58 size, and aging degree. This perspective aims to seek an *in situ* mass spectrometry  
59 imaging (MSI) method for MPs in organisms without labeling, and to quantitatively  
60 analyze plastics in imaging. This will deepen our understanding of the metabolic and  
61 transport mechanisms of MPs in biological tissues, and help us to evaluate the potential  
62 risks of MPs to organisms and even ecosystems.

## 63 2. Imaging characterization of MPs

64 Currently, microscopy imaging techniques are conventional detection methods for  
65 characterizing MPs within organisms (Table 1). Imaging techniques can be used to  
66 describe the physical properties (shape and size) of MPs and to quantify the observed  
67 MPs. In previous studies, laser confocal microscopy (CLSM), scanning electron  
68 microscopy (SEM), and atomic force microscopy (AFM) were used to observe

69 fluorescently labeled polystyrene (PS) MPs with a diameter of 10  $\mu\text{m}$  in the digestive  
 70 tracts of rotifers [20]. SEM was used to characterize the digestion of polylactic acid (PLA)  
 71 with a diameter of 25  $\mu\text{m}$  by gastric lipases in mice, and fluorescence microscopy (FM)  
 72 was used to characterize the migration process of fluorescent PLA plastic polymers in  
 73 mice [21]. Although these techniques are typically used to image artificially added  
 74 microsphere standards under laboratory conditions, they cannot specifically identify  
 75 environmental MPs in biological samples.

76

77 Table 1 Applications and limitations of conventional imaging techniques for MPs

Imaging techniques	Applications	Limitations	Ref.
Stereo microscope	<ul style="list-style-type: none"> <li>• Characterizing the size and morphology of MPs.</li> <li>• Counting the quantity of MPs.</li> <li>• Magnification to a certain degree to observe the details of MPs.</li> </ul>	<ul style="list-style-type: none"> <li>• Low magnification.</li> <li>• Unable to provide qualitative analysis, prone to false positives.</li> <li>• Not feasible for automation, time-consuming and labor-intensive.</li> <li>• Unable to perform <i>in situ</i> imaging of MPs.</li> </ul>	[13, 20]
FM	<ul style="list-style-type: none"> <li>• Characterizing the size and morphology of MPs.</li> <li>• Characterizing the migration and fragmentation behavior of MPs in environmental and biological samples.</li> </ul>	<ul style="list-style-type: none"> <li>• Fluorescent labeling of plastics is required, which may lead to false positive results due to dye leaching.</li> <li>• Fluorescence quenching can result in missed detection of MPs.</li> <li>• Chemical additives in synthetic plastics may exhibit fluorescence, interfering with the identification of MPs.</li> <li>• Environmental background may obscure the fluorescent signal of plastics.</li> </ul>	[6, 7, 20, 21, 43]
SEM/TEM	<ul style="list-style-type: none"> <li>• Characterizing high-resolution surface morphology of MPs.</li> <li>• Combining spectroscopic analysis techniques for chemical composition analysis of MPs.</li> <li>• Characterizing the distribution of MPs in environmental and biological samples.</li> </ul>	<ul style="list-style-type: none"> <li>• Coupled with energy-dispersive X-ray spectroscopy, the morphology and elemental composition of MPs can be determined. However, the specificity in identifying plastics is limited due to their common composition of C, H, and O.</li> <li>• Plastics cannot be quantitatively detected.</li> </ul>	[6, 8, 20, 21]
AFM	<ul style="list-style-type: none"> <li>• Surface imaging of nanoplastics.</li> <li>• Testing the mechanical properties of MPs.</li> <li>• Application of MP detection in environmental and</li> </ul>	<ul style="list-style-type: none"> <li>• The specificity in identifying plastics is limited.</li> <li>• Sample preparation requires a high level of precision, with the need for smooth and clean surfaces. Flexible or irregular plastic samples</li> </ul>	[20]

	biological samples.	are difficult to meet the requirements.	
Infrared spectroscopy imaging	<ul style="list-style-type: none"> <li>• Possessing micrometer-level spatial resolution.</li> <li>• No damage to the samples.</li> <li>• Identifying the chemical composition of MPs.</li> <li>• Rapid analysis with short detection time.</li> <li>• No need for additives or fluorescent labeling of MPs.</li> </ul>	<ul style="list-style-type: none"> <li>• Susceptible to environmental influences, such as temperature and humidity.</li> <li>• Plastics cannot be quantitatively detected.</li> <li>• High sample preparation requirements, unable to perform <i>in situ</i> imaging of plastics.</li> <li>• Plastics cannot be quantitatively detected.</li> </ul>	[44]
Near-infrared hyperspectral imaging	<ul style="list-style-type: none"> <li>• Possessing nanometer-level spatial resolution.</li> <li>• <i>In situ</i> imaging of MPs in environmental and biological samples without causing damage to the sample.</li> <li>• No need for additives or fluorescent labeling of MPs.</li> </ul>	<ul style="list-style-type: none"> <li>• Weak spectral specificity, prone to false positive results.</li> <li>• Limited ability to differentiate between plastics of different compositions.</li> <li>• Plastics cannot be quantitatively detected.</li> </ul>	[20, 43]
Raman spectroscopy imaging	<ul style="list-style-type: none"> <li>• Possessing nanometer-level spatial resolution.</li> <li>• Imaging of MPs without causing damage to the sample.</li> <li>• No need for additives or fluorescent labeling of MPs.</li> </ul>	<ul style="list-style-type: none"> <li>• Fluorescent substances in the sample (environmental matrix or plastic additives) can interfere with Raman signals.</li> <li>• Plastics cannot be quantitatively detected.</li> </ul>	[19, 20]

78

79 Compared to the above imaging techniques, vibrational spectroscopy imaging  
80 techniques, such as infrared spectroscopy imaging, near-infrared hyperspectral imaging,  
81 and Raman spectroscopy imaging, rely on the characteristic absorption peaks of plastics  
82 in their spectra for relatively precise identification of MPs. However, they are prone to  
83 false positive results due to matrix interference. These techniques require cumbersome  
84 matrix purification, and the MPs should be extracted before detection [20, 22]. This makes  
85 *in situ* imaging difficult to achieve.

86 Additionally, the imaging techniques have limitations in quantitative detection. It  
87 is challenging to identify particles smaller than the resolution of the instruments or  
88 particles firmly bound to biological matrices using imaging techniques. Although FM  
89 [22] and  $^{14}\text{C}$  isotope tracing techniques [23] can quantify the fluorescently or radioactively  
90 labeled plastic particles, their application in natural biological samples is difficult due

91 to the potential harm of labeled fluorescence and radioactivity to organisms, as well as  
92 their degradation and shedding under environmental conditions.

### 93 **3. MS detection techniques for MPs**

94 MS detection is widely used in the detection of trace organic pollutants in the  
95 environment. MS has the advantages of high sensitivity, high selectivity, high resolution,  
96 and fast detection speed, and can be used for both qualitative and quantitative analysis.  
97 Some MS techniques have been successfully applied in the detection of MPs or have  
98 potential applications for MP detection. Currently, plastic macromolecules need to be  
99 cracked or depolymerized into small molecules under high temperature, strong acid, or  
100 strong alkali conditions before MS detection [24-28]. This process cannot meet the  
101 demand for non-destructive imaging. However, there are some MS ionization methods  
102 that can achieve the direct ionization of plastic macromolecules.

#### 103 **3.1 Ionization of MP molecules**

104 The ionization of target analyte molecules is a prerequisite for MS detection.  
105 Gentle ionization methods, such as ambient ionization (e.g., Desorption Electrospray  
106 Ionization (DESI)), may struggle to ionize the plastic polymers with a high molecular  
107 weight of  $10^4$  Da. Therefore, vacuum ionization methods with higher ionization  
108 efficiency, such as matrix-assisted laser desorption/ionization (MALDI) MS and  
109 secondary ion mass spectrometry (SIMS), hold greater potential for the detection of  
110 plastic polymers [29-31].

111 SIMS is a hard-ionization technique that can bombard macromolecules into  
112 complex fragments, making it suitable for detecting plastic polymers [32]. SIMS  
113 bombards the sample surface with high-energy particles (nitrogen or argon), causing  
114 the atoms and molecules on the sample surface to become ionized. After bombardment,  
115 secondary ions are produced from atoms and molecules, which then enter the mass  
116 analyzer for analysis [32]. The spatial resolution of this technique reaches the nanometer  
117 level. However, the extensive ion fragmentation and low sensitivity limit the  
118 application of SIMS in trace pollutants imaging (Table S1).

119 MALDI is an ion source that utilizes laser energy to vaporize the matrix with the  
120 aid of ionization reagents, resulting in the ionization of target analyte molecules. The



121 matrix is crucial because it should possess the property of absorbing laser energy and  
122 converting it into thermal energy. Ionization reagents serve to enhance the ionization  
123 efficiency of the sample and increase the intensity of the mass spectrum signal. After  
124 mixing the sample, matrix, and ionization reagents together, a thin film forms on the  
125 surface of the sample. Subsequently, the laser irradiates the surface of the sample matrix,  
126 causing the matrix molecules to evaporate while absorbing the laser energy. As a result,  
127 the sample is released and ionized. The ions of the molecular fragmentation products  
128 then enter the mass analyzer for analysis and detection. After absorbing the laser energy,  
129 the matrix undergoes dissociation or fragmentation processes, resulting in the  
130 generation of charged ions. During this process, interactions occur between matrix ions  
131 and analyte molecules, facilitating proton transfer to protonate the analyte. The  
132 assistance of matrix and ionization reagents significantly improves the ionization  
133 efficiency of the sample, addressing the issue of ionizing non-volatile and high-  
134 molecular-weight analytes in MS. In contrast to the ESI techniques such as DESI, which  
135 are suitable for ionizing water-soluble compounds with molecular weights below 2000  
136 Da, MALDI ionization is performed under vacuum conditions. This makes MALDI  
137 ionization more appropriate for ionizing lipophilic macromolecular polymers with  
138 molecular weights of tens of thousands of Da. MALDI-time-of-flight mass  
139 spectrometry (MALDI-TOF-MS) has been successfully applied in the detection of MPs.  
140 Professor Cai's team quantified PS and polyethylene terephthalate (PET) in sediments  
141 and aviation plastic cups using MALDI-TOF-MS<sup>[33]</sup>. The spatial resolution of MALDI-  
142 TOF can reach up to 1-5  $\mu\text{m}$ , which offers great potential advantages in imaging  
143 detection (Table S1).

### 144 **3.2 Mass analysis of MP molecules**

145 The MS detection of polymers also imposes requirements on the mass analyzer.  
146 Mass analyzers<sup>[24]</sup> such as TOF, Orbitrap, and magnetic sector MS can be used for the  
147 detection of biomacromolecules, and theoretically, they can also be applied to the  
148 detection of MPs.

149 TOF-MS is renowned for its capability in analyzing macromolecules. Ionized  
150 molecules are accelerated by an electric field and fly at a constant velocity in the flight

151 tube, unaffected by external forces. Due to the variance in ion mass, their flight  
152 velocities differ, leading to varied arrival times at the detector. The mass can be  
153 determined based on the flight time. Therefore, TOF mass analyzers have  $m/z$ -  
154 independent trapping conditions [24]. They also have other advantages such as the  
155 highest detection speed and high resolution. However, the susceptibility to  
156 environmental influences during the detection process, to a certain extent, limits the  
157 application of TOF in terms of mass accuracy, resolution, and sensitivity (Table S1).

158 Orbitrap is a modified ion trap mass analyzer. After entering the orbitrap, ions  
159 undergo radial motion under the influence of an electric field and axial oscillation. Due  
160 to variance in ion mass, ions exhibit different frequencies in the orbit. The accumulation  
161 of charge within the orbit generates a current signal outside the orbitrap, which is  
162 detected and converted into a mass spectrum [34, 35]. The orbitrap mass analyzer has high  
163 resolving power, high mass accuracy, and short acquisition times. Theoretically,  
164 orbitrap has an unlimited mass range [36] and can be used for the analysis of  
165 macromolecules. Compared with the orbitrap, Fourier transform ion cyclotron  
166 resonance mass spectrometry (FTICR-MS) has higher trapping efficiencies for large  
167 ions, increased isotopic fidelity, and more precise resolution (Table S1). However, the  
168 high accuracy of FTICR MS comes at the cost of longer signal acquisition times. Mass  
169 imaging detection requires frame-by-frame information acquisition for a sample, and  
170 the precision of a frame is usually 20-100 microns. Thus, an excessively long  
171 acquisition time of tens of hours will be needed for one sample. The relative standard  
172 deviation (RSD) of both TOF and Orbitrap in environmental samples can generally be  
173 achieved to be less than 3% (Table S1). In general, TOF or Orbitrap is sufficient for  
174 analyzing high-molecular-weight compounds.

175 The magnetic sector MS is also a mass analyzer capable of analyzing  
176 macromolecules. The ionized sample, under the influence of a magnetic field,  
177 experiences Lorentz force and deflects along a trajectory. Due to variance in ion mass,  
178 ions are separated in mass-to-charge ratio orbits. Then, the separated ions are recorded  
179 by the detector. Isotope ratio mass spectrometry (IRMS), as a type of magnetic sector  
180 MS, has been used to trace the sources of MPs rather than their abundance [24]. However,

181 IRMS requires complete decomposition of organic compounds at temperatures  
182 exceeding 1,000 °C to ensure accurate isotope testing results [37]. Due to the breakdown  
183 of plastic macromolecules, this type of MS is unsuitable for the analysis of MPs.

#### 184 **4. Prospect of *in situ* imaging of MPs in biological samples using MSI**

185 Scientists are seeking a detection technology that can provide *in situ* spatial  
186 distribution information of samples, as well as the quality and quantity of MPs. MSI  
187 combines microscopic imaging and MS, enabling the acquisition of both surface  
188 morphology in samples and mass spectra information of various chemical compounds.  
189 By matching and overlaying these two types of information, high-resolution images  
190 with chemical composition analysis can be obtained.

191 MSI is an MS technique that evolved from measuring the spatial distribution of  
192 endogenous compound molecules in biological tissues [38]. In 1998, the distribution  
193 characteristics of phospholipids on cell membranes were studied using TOF-SIMS [39].  
194 Recently, MSI has been used to characterize the distribution of exogenous  
195 environmental pollutants within organisms. For example, the distribution of  
196 imipramine and chloroquine in the kidneys and brains of mice was visualized using  
197 atmospheric pressure-MALDI-TOF [40].

198 The development of *in situ* imaging and quantitative methods for MPs will help  
199 understand the migration and transformation processes of MPs in the environment, as  
200 well as the transport mechanisms in biological tissues. MALDI and SIMS coupled with  
201 TOF or Orbitrap mass analyzers are proficient in analyzing and identifying  
202 macromolecules such as plastic polymers. The reported concentration of PET MPs in  
203 mussels from the market in Tianjin is 75.4 ng/g, while 12 MP particles with sizes of 5  
204 to 10 µm were detected in four human placentas [13]. According to Hermabessiere's  
205 calculation method [25], the mass concentration of polypropylene (PP) plastic is  
206 estimated to be 10<sup>3</sup> ng/g from the quantity concentration. This is significantly higher  
207 than the Instrumental Detection Limits (IDLs) for MALDI-TOF-MS detection, which  
208 are 5.2 ng [27,32]. By combining *in situ* ionization techniques such as MALDI and SIMS  
209 with MS like TOF or Orbitrap (which are proficient in macromolecular analysis and  
210 identification), this technology will perform excellently in the analysis and

211 identification of plastic MPs. By combining the collected MS data with the spatial  
212 information of optical images, the spatial images of chemical compositions can be  
213 generated, theoretically enabling the characterization of plastic polymers through MSI  
214 (Fig. 1). Attempts to perform plastic polymer *in situ* imaging using MALDI-TOF-MS<sup>[41]</sup>  
215 and TOF-SIMS <sup>[42]</sup> have begun (Fig. 2).

216

217 **Fig. 1** Schematic diagram of an ideal MSI technique.

218

219 **Fig. 2** Optical image (a), surface layer-MALDI-MS images (b-d) of a bilayer film  
220 prepared with PMMA (green) and PS (purple), intensity profile (e) along the red line in  
221 (d) <sup>[41]</sup>. Optical image (f-j) and surface layer-MALDI-MS images (g-k) of logo patterns  
222 scribed on the surface of PMMA film <sup>[41]</sup>. Simultaneous visualization of three different  
223 types of MPs (PMMA, GMA, and PVC) from a paramecium (l-p) <sup>[42]</sup>.

224

225 Due to the hard ionization principle of TOF-SIMS, plastic macromolecules can be  
226 fragmented into complex fragments upon bombardment. The resulting mass spectral  
227 signals may not be sufficiently clear, leading to inaccurate quantification <sup>[38]</sup>. In contrast,  
228 MALDI-MSI, based on soft ionization principles, can ionize plastic macromolecules  
229 without disrupting their molecular structure. Thus, MALDI-MSI holds potential for the  
230 accurate quantification of MPs <sup>[33]</sup>. However, several challenges remain in establishing  
231 methods for detecting MPs using MALDI-MSI.

232 1) How to remove interference from embedding agents: The difficulty in preparing  
233 biological samples for slicing without embedding agents. Embedding agents used to  
234 support the slicing of biological tissues may produce background interference. If the  
235 embedding material is necessary, its background noise should be identified by carefully  
236 comparing it with the characteristic peaks of the target molecules. Cryosectioning  
237 without embedding is the best choice, which requires finding the optimal slicing  
238 conditions (including adjusting the method and duration) for freezing the samples.  
239 Besides, some semi-synthetic plastics, such as Celluloid, which is made from cellulose,  
240 may have molecular structures and MS signals similar to those of plant biological

241 tissues. This may cause extra difficulties in conducting MS-based detection of such  
242 substances.

243 2) How to improve imaging speed. High resolution of the instrument and oversized  
244 sample areas can result in excessively long imaging times. Introducing machine  
245 learning and deep learning techniques can aid in developing more efficient and rapid  
246 MSI data processing programs.

247 3) How to improve the accuracy of MSI quantification. Inaccuracies in quantifying  
248 MPs may arise due to matrix interference and ion suppression in biological samples.  
249 Attempting internal standard correction for quantification or employing multiple  
250 calibration methods, such as thermal decomposition coupled with gas chromatography-  
251 MS or thermal alkaline/acid-assisted liquid chromatography-MS, can enhance the  
252 accuracy of quantification results from MSI.

253

#### 254 **CRedit authorship contribution statement**

255 Y.L.: writing—original draft, review, editing & funding acquisition; X.Y.S., Y.W.:  
256 writing—review & editing; Y.F. Z.: resources, formal analysis; J.J.Z.: formal analysis,  
257 validation, writing—review & editing; P.W.: formal analysis; X.F.C.: supervision,  
258 resources, validation; B.S.X.: supervision, writing; L.W.: methodology,  
259 conceptualization, supervision, writing—review & editing, funding acquisition.

260

#### 261 **Declaration of competing interest**

262 The authors declare that they have no known competing financial interests or personal  
263 relationships that could have appeared to influence the work reported in this paper.

264

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270

271 **Reference:**

- 272 [1] A.A. De Souza Machado, W. Kloas, C. Zarfl, S. Hempel, M.C. Rillig, Microplastics as an  
273 emerging threat to terrestrial ecosystems, *Global Change Biol.* 24 (2018) 1405-1416,  
274 <https://doi.org/10.1111/gcb.14020>.
- 275 [2] R.C. Thompson, Y. Olsen, R. P. Michell, A. Davis, S. J. Rowland, A.W.G. John, D. McGonigle,  
276 A.E. Russell, Lost at sea: where is all the plastics? *Science* 304 (2004) 838,  
277 <https://doi.org/10.1126/science.1094559>.
- 278 [3] J. Alexander, L. Barregard, M. Bignami, S. Ceccatelli, B. Cottrill, M. Dinovi, L. Edler, B. Grasl-  
279 Kraupp, C. Hogstrand, et al., Presence of microplastics and nanoplastics in food, with particular  
280 focus on seafood, *EFSA J.* 14 (2016) 4501, <https://doi.org/10.2903/j.efsa.2016.4501>
- 281 [4] H.K. Mcllwraith, J. Kim, P. Helm, S.P. Bhavsar, J.S. Metzger, C.M. Rochman, Evidence of  
282 microplastics translocation in wild-caught fish and implications for microplastic accumulation  
283 dynamics in food webs, *Environ. Sci. Technol.* 55 (2021) 12372-12382,  
284 <https://doi.org/10.1021/acs.est.1c02922>.
- 285 [5] F. Ribeiro, E.D. Okoffo, J.W. O'Brien, S. Fraissinet-Tachet, S. O'Brien, M. Gallen, S.  
286 Samanipour, S. Kaserzon, et al., Quantitative analysis of selected plastics in high-commercial-value  
287 Australian seafood by pyrolysis gas chromatography mass spectrometry, *Environ. Sci. Technol.* 54  
288 (2020) 9408-9417, <https://doi.org/10.1021/acs.est.0c02337>.
- 289 [6] X.D. Sun, X.Z. Yuan, Y. Jia, L.J. Feng, F.P. Zhu, S.S. Dong, J. Liu, X. Kong, et al., Differentially  
290 charged nanoplastics demonstrate distinct accumulation in *Arabidopsis thaliana*, *Nat. Nanotechnol.*  
291 15 (2020) 755-760, <https://doi.org/10.1038/s41565-020-0707-4>.
- 292 [7] L. Li, Y. Luo, R. Li, Q. Zhou, W.J.G.M. Peijnenburg, N. Yin, J. Yang, C. Tu, Y. Zhang, Effective  
293 uptake of submicrometre plastics by crop plants via a crack-entry mode, *Nat. Sustain.* 3 (2020) 929-  
294 937, <https://doi.org/10.1038/s41893-020-0567-9>.
- 295 [8] Y. Luo, L. Li, Y. Feng, R. Li, J. Yang, W. Peijnenburg, C. Tu, Quantitative tracing of uptake and  
296 transport of submicrometre plastics in crop plants using lanthanide chelates as a dual-functional  
297 tracer, *Nat. Nanotechnol.* 17 (2022) 424-431, <https://doi.org/10.1038/s41565-021-01063-3>.
- 298 [9] L. Zhu, M. Ma, X. Sun, Z. Wu, Y. Yu, Y. Kang, Z. Liu, Q. Xu, L. An, Microplastics entry into  
299 the blood by infusion therapy: few but a direct pathway, *Environ. Sci. Technol. Lett.* 11 (2023) 67-  
300 72, <https://doi.org/10.1021/acs.estlett.3c00905>.

- 301 [10] H.A. Leslie, M.J.M. van Velzen, S.H. Brandsma, A.D. Vethaak, J.J. Garcia-Vallejo, M.H.  
302 Lamoree, Discovery and quantification of plastic particle pollution in human blood, *Environ. Int.*  
303 163 (2022) 107199, <https://doi.org/10.1016/j.envint.2022.107199>.
- 304 [11] K. Mattsson, E.V. Johnson, A. Malmendal, S. Linse, L.A. Hansson, T. Cedervall, Brain damage  
305 and behavioural disorders in fish induced by plastic nanoparticles delivered through the food chain,  
306 *Sci. Rep.* 7 (2017) 11452, <https://doi.org/10.1038/s41598-017-10813-0>.
- 307 [12] M. Prust, J. Meijer, R.H.S. Westerink, The plastic brain: neurotoxicity of micro- and  
308 nanoplastics, *Part. Fibre Toxicol.* 17 (2020) 24, <https://doi.org/10.1186/s12989-020-00358-y>.
- 309 [13] A. Ragusa, A. Svelato, C. Santacroce, P. Catalano, V. Notarstefano, O. Carnevali, F. Papa,  
310 M.C.A. Rongioletti, F. Baiocco, S. Draghi, E. D'Amore, D. Rinaldo, M. Matta, E. Giorgini,  
311 *Plasticenta: First evidence of microplastics in human placenta*, *Environ. Int.* 146 (2021) 106274,  
312 <https://doi.org/10.1016/j.envint.2020.106274>.
- 313 [14] S. Grafmueller, P. Manser, L. Diener, P.A. Diener, X. Maeder-Althaus, L. Maurizi, W. Jochum,  
314 H.F. Krug, T. Buerki-Thurnherr, U. von Mandach, P. Wick, Bidirectional transfer study of  
315 polystyrene nanoparticles across the placental barrier in an ex vivo human placental perfusion model,  
316 *Environ. Health Perspect.* 123 (2015) 1280-1286, <https://doi.org/10.1289/ehp.1409271>.
- 317 [15] X. Yang, Y.B. Man, M.H. Wong, R.B. Owen, K.L. Chow, Environmental health impacts of  
318 microplastics exposure on structural organization levels in the human body, *Sci. Total Environ.* 825  
319 (2022) 154025, <https://doi.org/10.1016/j.scitotenv.2022.154025>.
- 320 [16] L.J. Mortensen, G. Oberdorster, A.P. Pentland, L.A. Delouise, In vivo skin penetration of  
321 quantum dot nanoparticles in the murine model: the effect of UVR, *Nano Lett.* 8 (2008) 2779-2787,  
322 <https://doi.org/10.1021/nl801323y>.
- 323 [17] I. Fiorentino, R. Gualtieri, V. Barbato, V. Mollo, S. Braun, A. Angrisani, M. Turano, M. Furia,  
324 P.A. Netti, D. Guarnieri, S. Fusco, R. Talevi, Energy independent uptake and release of polystyrene  
325 nanoparticles in primary mammalian cell cultures, *Exp. Cell Res.* 330 (2015) 240-247,  
326 <https://doi.org/10.1016/j.yexcr.2014.09.017>.
- 327 [18] J. Lee, K.J. Chae, A systematic protocol of microplastics analysis from their identification to  
328 quantification in water environment: a comprehensive review, *J. Hazard. Mater.* 403 (2021) 124049,  
329 <https://doi.org/10.1016/j.jhazmat.2020.124049>.
- 330 [19] J. Zhang, M. Peng, E. Lian, L. Xia, A.G. Asimakopoulos, S. Luo, L. Wang, Identification of



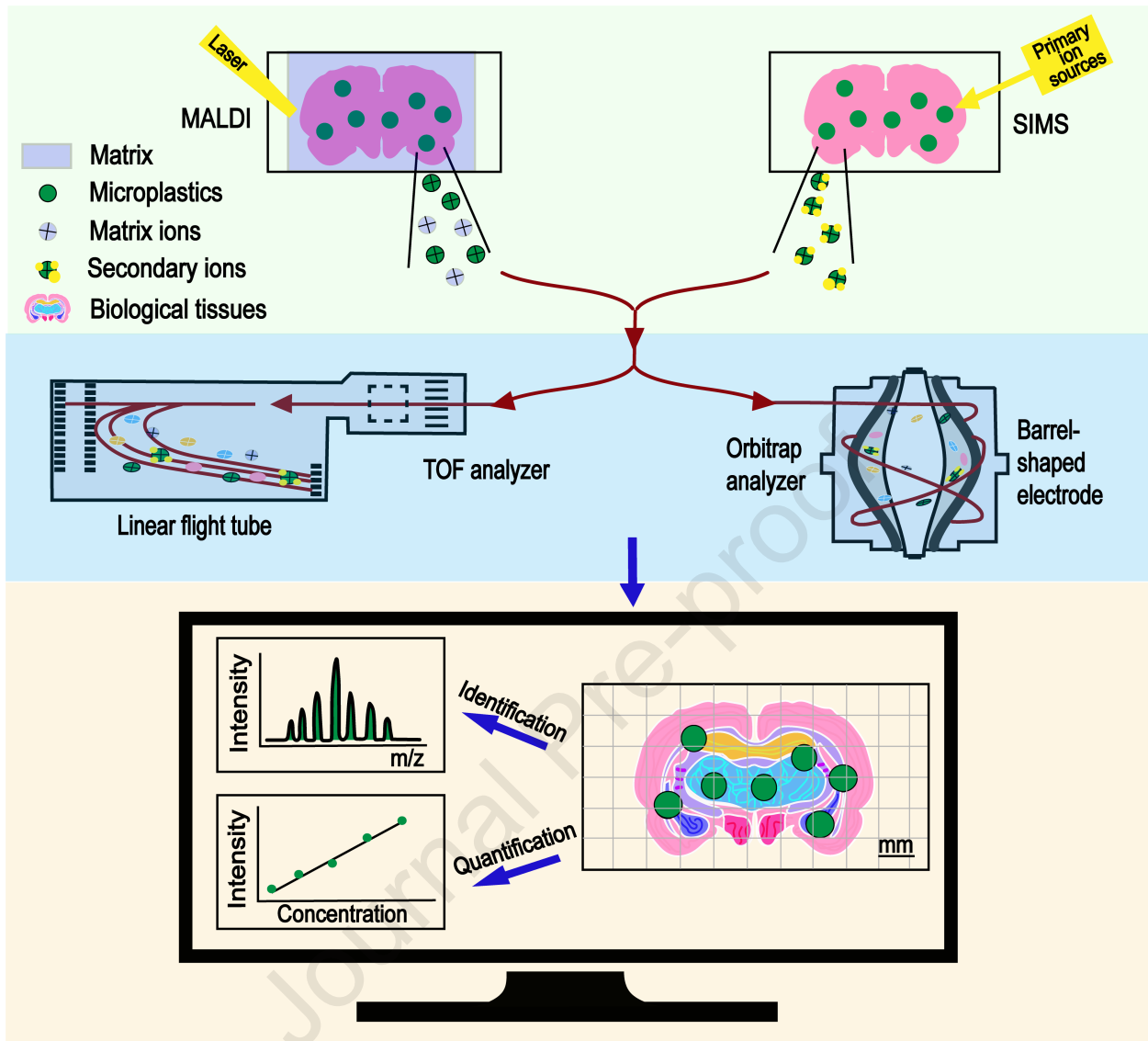
- 331 poly(ethylene terephthalate) nanoplastics in commercially bottled drinking water using surface-  
332 enhanced Raman spectroscopy, *Environ. Sci. Technol.* 57 (2023) 8365-8372,  
333 <https://doi.org/10.1021/acs.est.3c00842>.
- 334 [20] J. Zhao, R. Lan, Z. Wang, W. Su, D. Song, R. Xue, Z. Liu, X. Liu, Y. Dai, T. Yue, B. Xing,  
335 Microplastic fragmentation by rotifers in aquatic ecosystems contributes to global nanoplastic  
336 pollution, *Nat. Nanotechnol.* (2023), <https://doi.org/10.1038/s41565-023-01534-9>.
- 337 [21] M. Wang, Q. Li, C. Shi, J. Lv, Y. Xu, J. Yang, S.L. Chua, L. Jia, H. Chen, Q. Liu, C. Huang, Y.  
338 Huang, J. Chen, M. Fang, Oligomer nanoparticle release from polylactic acid plastics catalysed by  
339 gut enzymes triggers acute inflammation, *Nat. Nanotechnol.* 18 (2023) 403-411,  
340 <https://doi.org/10.1038/s41565-023-01329-y>.
- 341 [22] S. Rist, A. Baun, N.B. Hartmann, Ingestion of micro- and nanoplastics in *Daphnia magna*-  
342 Quantification of body burdens and assessment of feeding rates and reproduction, *Environ. Pollut.*  
343 228 (2017) 398-407, <https://doi.org/10.1016/j.envpol.2017.05.048>.
- 344 [23] L. Tian, Y. Ma, R. Ji, Quantification of polystyrene plastics degradation using  $^{14}\text{C}$  isotope tracer  
345 technique, *Methods Enzymol.* 648 (2021) 121-136, <https://doi.org/10.1016/bs.mie.2020.12.014>.
- 346 [24] J. Zhang, D. Fu, H. Feng, Y. Li, S. Zhang, C. Peng, Y. Wang, H. Sun, L. Wang, Mass  
347 spectrometry detection of environmental microplastics: advances and challenges, *TrAC Anal. Chem.*  
348 170 (2024) 117472, <https://doi.org/10.1016/j.trac.2023.117472>.
- 349 [25] L. Hermabessiere, C. Himber, B. Boricaud, M. Kazour, R. Amara, A.L. Cassone, M. Laurentie,  
350 I. Paul-Pont, P. Soudant, A. Dehaut, G. Duflos, Optimization, performance, and application of a  
351 pyrolysis-GC/MS method for the identification of microplastics, *Anal. Bioanal. Chem.* 410 (2018)  
352 6663-6676, <https://doi.org/10.1007/s00216-018-1279-0>.
- 353 [26] L. Wang, Y. Peng, Y. Xu, J. Zhang, T. Zhang, M. Yan, H. Sun, An *in situ* depolymerization and  
354 liquid chromatography-tandem mass spectrometry method for quantifying polylactic acid  
355 microplastics in environmental samples, *Environ. Sci. Technol.* 56 (2022) 13029-13035,  
356 <https://doi.org/10.1021/acs.est.2c02221>.
- 357 [27] L. Wang, J. Zhang, S. Hou, H. Sun, A simple method for quantifying polycarbonate and  
358 polyethylene terephthalate microplastics in environmental samples by liquid chromatography-  
359 tandem mass spectrometry, *Environ. Sci. Technol. Lett.* 4 (2017) 530-534,  
360 <https://doi.org/10.1021/acs.estlett.7b00454>.

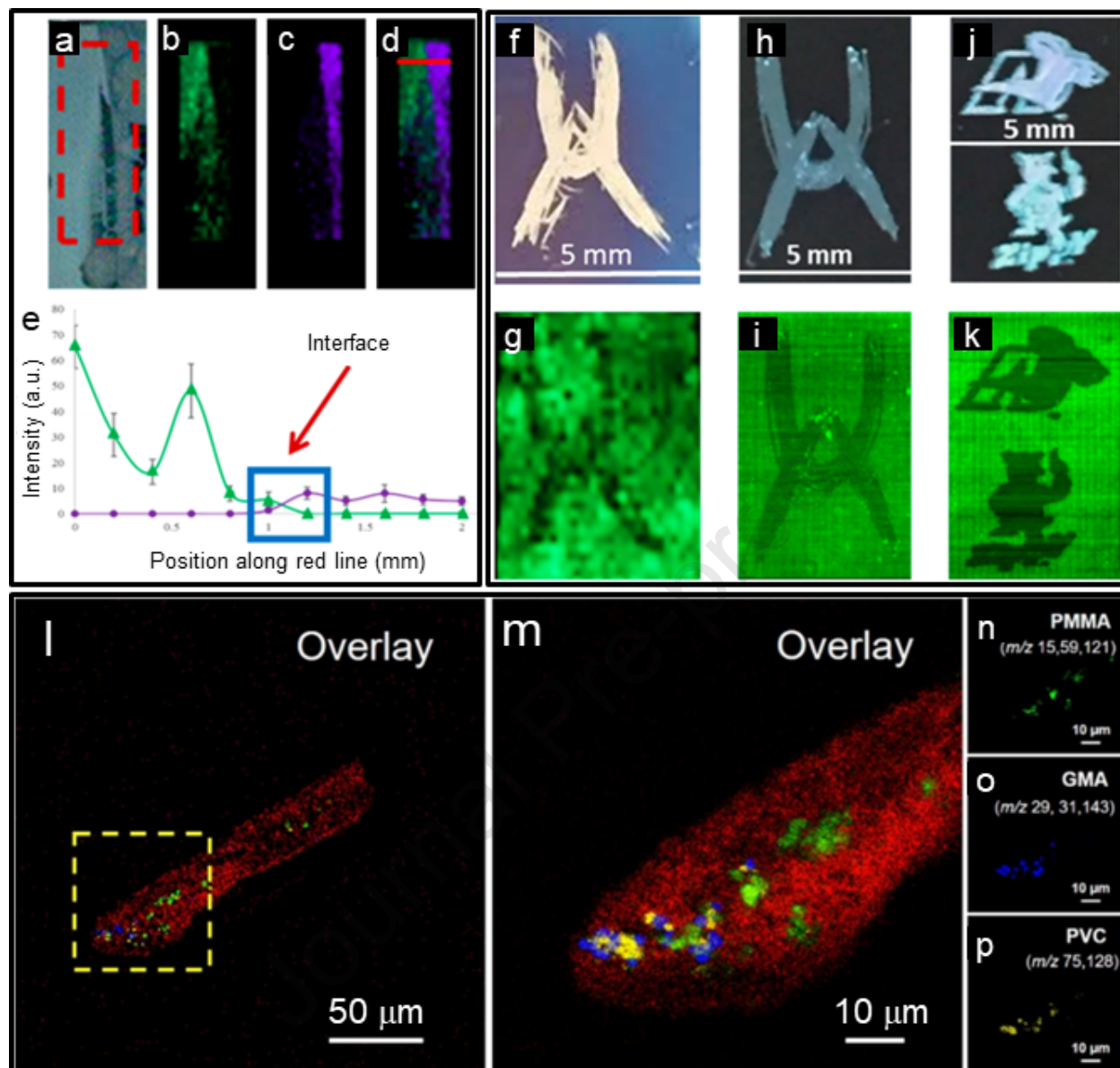


- 361 [28] C. Peng, X. Tang, X. Gong, Y. Dai, H. Sun, L. Wang, Development and application of a mass  
362 spectrometry method for quantifying nylon microplastics in environment, *Anal. Chem.* 92 (2020)  
363 13930-13935, <https://doi.org/10.1021/acs.analchem.0c02801>.
- 364 [29] C. Wu, A.L. Dill, L.S. Eberlin, R.G. Cooks, D.R. Ifa, Mass spectrometry imaging under ambient  
365 conditions, *Mass Spec. Rev.* 32 (2013) 218-243, <https://doi.org/10.1002/mas.21360>.
- 366 [30] Y.E. Kim, S.Y. Yi, C.S. Lee, Y. Jung, B.H. Chung, Gold patterned biochips for on-chip immuno-  
367 MALDI-TOF MS: SPR imaging coupled multi-protein MS analysis, *Analyst* 137 (2012) 386-392,  
368 <https://doi.org/10.1039/c1an15659d>.
- 369 [31] P. Pompach, O. Benada, M. Rosulek, P. Darebna, J. Hausner, V. Ruzicka, M. Volny, P. Novak,  
370 Protein chips compatible with MALDI mass spectrometry prepared by ambient ion landing, *Anal.*  
371 *Chem.* 88 (2016) 8526-8534, <https://doi.org/10.1021/acs.analchem.6b01366>.
- 372 [32] C. Bich, D. Touboul, A. Brunelle, Biomedical studies by TOF-SIMS imaging, *Biointerphases*  
373 10 (2014) 018901, <https://doi.org/10.1116/1.4901511>.
- 374 [33] P. Wu, Y. Tang, G. Cao, J. Li, S. Wang, X. Chang, M. Dang, H. Jin, C. Zheng, Z. Cai,  
375 Determination of environmental micro(nano)plastics by matrix-assisted laser desorption/ionization-  
376 time-of-flight mass spectrometry, *Anal. Chem.* 92 (2020) 14346-14356,  
377 <https://doi.org/10.1021/acs.analchem.0c01928>.
- 378 [34] R.A. Zubarev, A. Makarov, Orbitrap mass spectrometry, *Anal. Chem.* 85 (2013) 5288-5296,  
379 <https://doi.org/10.1021/ac4001223>.
- 380 [35] M. Llorca, A. Vega-Herrera, G. Schirinzi, K. Savva, E. Abad, M. Farre, Screening of suspected  
381 micro(nano)plastics in the Ebro Delta (Mediterranean Sea), *J. Hazard. Mater.* 404 (2021) 124022,  
382 <https://doi.org/10.1016/j.jhazmat.2020.124022>.
- 383 [36] J.B. Shaw, J.S. Brodbelt, Extending the isotopically resolved mass range of Orbitrap mass  
384 spectrometers, *Anal. Chem.* 85 (2013) 8313-8318, <https://doi.org/10.1021/ac401634b>.
- 385 [37] Q.T. Birch, P.M. Potter, P.X. Pinto, D.D. Dionysiou, S.R. Al-Abed, Isotope ratio mass  
386 spectrometry and spectroscopic techniques for microplastics characterization, *Talanta* 224 (2021)  
387 121743, <https://doi.org/10.1016/j.talanta.2020.121743>.
- 388 [38] B.A. Boughton, D. Thinakaran, D. Sarabia, A. Bacic, U. Roessner, Mass spectrometry imaging  
389 for plant biology: a review, *Phytochem. Rev.* 15 (2016) 445-488, [https://doi.org/10.1007/s11101-](https://doi.org/10.1007/s11101-015-9440-2)  
390 015-9440-2.

- 391 [39] M.L. Pacholski, D.M. Cannon, A.G. Ewing, N. Winograd, Static time-of-flight secondary ion  
392 mass spectrometry imaging of freeze-fractured, frozen-hydrated biological membranes, *Rapid*  
393 *Commun. Mass Spectrom.* 12 (1998) 1232-1235, [https://doi.org/10.1002/\(SICI\)1097-](https://doi.org/10.1002/(SICI)1097-0231(19980930)12:183.3.CO;2-7)  
394 [0231\(19980930\)12:183.3.CO;2-7](https://doi.org/10.1002/(SICI)1097-0231(19980930)12:183.3.CO;2-7).
- 395 [40] A. Islam, T. Sakamoto, Q. Zhai, M.M. Rahman, M.A. Mamun, Y. Takahashi, T. Kahyo, M.  
396 Setou, Application of AP-MALDI imaging mass microscope for the rapid mapping of imipramine,  
397 chloroquine, and their metabolites in the kidney and brain of wild-type mice, *Pharmaceuticals* 15  
398 (2022) 1314, <https://doi.org/10.3390/ph15111314>.
- 399 [41] K.J. Endres, J.A. Hill, K. Lu, M.D. Foster, C. Wesdemiotis, Surface layer matrix-assisted laser  
400 desorption ionization mass spectrometry imaging: a surface imaging technique for the molecular-  
401 level analysis of synthetic material surfaces, *Anal. Chem.* 90 (2018) 13427-13433,  
402 <https://doi.org/10.1021/acs.analchem.8b03238>.
- 403 [42] J. Feng, H. Zhao, X. Gong, M.C. Xia, L. Cai, H. Yao, X. Zhao, Z. Yan, Z. Li, H. Nie, X. Ma,  
404 S. Zhang, *In situ* identification and spatial mapping of microplastic standards in paramecia by  
405 secondary-ion mass spectrometry imaging, *Anal. Chem.* 93 (2021) 5521-5528,  
406 <https://doi.org/10.1021/acs.analchem.0c05383>.
- 407 [43] R. Xue, R. Lan, W. Su, Z. Wang, X. Li, J. Zhao, C. Ma, B. Xing, Mechanistic understanding  
408 toward the maternal transfer of nanoplastics in *Daphnia magna*, *ACS Nano* 17 (2023) 13488-13499,  
409 <https://doi.org/10.1021/acsnano.3c01847>.
- 410 [44] Y. Tian, Z. Chen, J. Zhang, Z. Wang, Y. Zhu, P. Wang, T. Zhang, J. Pu, H. Sun, L. Wang, An  
411 innovative evaluation method based on polymer mass detection to evaluate the contribution of  
412 microfibers from laundry process to municipal wastewater, *J. Hazard. Mater.* 407 (2021) 124861,  
413 <https://doi.org/10.1016/j.jhazmat.2020.124861>.

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### **Highlights**

- Mass spectrometry imaging holds promise for in situ imaging of microplastics in biological samples.
- Mass spectrometry imaging could provide in situ spatial distribution information of biological samples
- Mass spectrometry imaging could quantify the quality and quantity of MPs in biological samples.